

24 efficient use of water. However, information on the water-use of indigenous trees
25 and forests is scarce and indirect, and the relative contributions of transpiration,
26 canopy interception and litter interception to total evaporation have until now not
27 been investigated in South Africa. To quantify these fluxes, both field
28 measurements and modelling were undertaken. In this study, green water-use by
29 indigenous *Podocarpus henkelii* and an exotic species *Pinus patula* are
30 compared. The results from this study show that the productive green water-use
31 by *P. henkelii* and *P. patula* was 41.0% and 95.9% of gross precipitation
32 respectively over the 18 month period of this study. The non-productive canopy
33 and litter interception by *P. henkelii* accounts for 29.8% and 6.2% respectively,
34 while canopy and litter interception accounted for 22.1% and 10.7% respectively
35 for *P. patula*. The productive green WUE of *P. henkelii* and *P. patula* is
36 7.14g.mm^{-1} and 25.21g.mm^{-1} respectively, in comparison with the total green
37 WUE of 3.8g.mm^{-1} and 18.8g.mm^{-1} . From a water resources management and
38 planning perspective it is important to consider the total green WUE, but also to
39 have a good understanding of the relative contributions of each component of the
40 green water fluxes so that water abstracted from the soil can be differentiated
41 from the water that does not reach the soil due to losses of canopy and litter
42 interception and does not get lumped as one evaporative loss.

43

44 **Keywords:** Water-use efficiency; *Podocarpus henkelii*; *Pinus patula*;
45 transpiration; rainfall interception

46

47 **INTRODUCTION**

48

49 The production of biomass for direct human consumption as food and timber is
50 by far the largest human managed consumer of freshwater on Earth (Falkenmark
51 and Rockström, 2006). The limited supply of indigenous timber means South
52 Africa has a large area of exotic forest plantations that are planted in the wetter
53 regions of the country, covering an area of approximately 1.4 million hectares
54 compared with 0.5 million hectares covered by indigenous forests. The
55 commercial forestry sector contributes approximately 22 billion Rand to the South
56 Africa economy and employs approximately 170 000 people (Chamberlain *et al.*,
57 2005; DAFF, 2010). The “Green Water” concept where all vapour fluxes including
58 transpiration, soil evaporation, canopy interception and litter interception are
59 considered was introduced by Falkenmark (1995) and has since gained
60 prominence as a highly effective way of highlighting the role of evaporation from
61 the landscape (Jewitt, 2006). The consideration of green water flows in formal
62 water resources planning is however proving to be very difficult (Jewitt, 2006),
63 not least because of confusion in terminology.

64

65 There have been a number of studies undertaken in South Africa to quantify the
66 green water-use (total evaporation) of introduced commercial forestry species.
67 These studies have shown conclusively that green water-use from commercial
68 forest plantations is substantially higher than from the original grasslands or
69 fynbos that were replaced by afforestation and that this results in reduced

70 streamflows (Blue water) (Dye, 1996; Scott *et al.*, 2000). Thus, forest plantations
71 have mostly reduced catchment water yields, and this has resulted in legislation
72 limiting further afforestation in areas where water supplies are already
73 committed. Although the demand for timber is growing strongly, the extent of the
74 national forestry estate is essentially capped to minimise further declines in
75 surface “Blue Water” resources (Dye *et al.*, 2008). The production of biomass is a
76 water-dependant process. During photosynthesis, when the stomata are open to
77 take in carbon dioxide, a large amount of water is simultaneously being
78 transpired. While transpiration is considered a productive green water flow as it is
79 responsible for biomass production, it is accompanied by non-productive
80 evaporative losses from the soil, litter and canopy should water be available.
81 Together, these vapour fluxes of transpiration, soil evaporation, and canopy and
82 litter interception constitute the total Green water used in biomass production
83 (Falkenmark and Rockström, 2006).

84

85 There is a widespread belief within South Africa that indigenous tree species, in
86 contrast to commercial forestry genera/species including *Pinus* (pine),
87 *Eucalyptus* (gum) and *Acacia mearnsii* (wattle), are “water-wise” and should be
88 planted more widely in view of their perceived more efficient use of water. This
89 perception appears to be based on the observation that indigenous trees are
90 generally slow growers, and the belief that growth rate and water-use are broadly
91 linked. However, tree water-use, and the total evaporation from forests and
92 woodlands, is difficult to measure, and so evidence of low water-use by

93 indigenous trees is scarce and indirect. The most comprehensive study of water-
94 use by indigenous trees in South Africa was undertaken by Dye *et al.*, (2008) and
95 largely forms the benchmark against which the results from this study are
96 compared. Typically, water-use efficiency (WUE) studies express water-use in
97 terms of an increase in wood biomass relative to transpiration (i.e. productive
98 Green water-use). In this study, only the stem biomass was considered, and did
99 not include the branches and leaves. From the findings of a study by Dye *et al.*,
100 (2008) which considered the productive Green WUE of indigenous species, it
101 was found that the WUE of indigenous species is generally lower than that of
102 commercial forestry species. Dye *et al.*, (2008) defined WUE in their study as the
103 increase in stem dry biomass per unit of water transpired. Although transpiration
104 by the indigenous species was generally lower in comparison to more productive
105 commercial species, particularly *Eucalyptus grandis*, the rate of growth was also
106 much slower, and result in a lower WUE. Dye *et al.*, (2008) concluded that in
107 general, indigenous trees appear to possess an advantage over commercial
108 species on productive sites in having lower water-use and lower streamflow
109 reduction impact, but not in growth rate. Because Dye *et al.*, (2008) based their
110 study on transpiration measurements only, the non-productive component of total
111 evaporation (i.e. interception) was not considered. Because of the scarcity of
112 studies on water-use of indigenous trees and forests the relative contributions of
113 transpiration, canopy interception and litter interception to total evaporation have,
114 until now, not been investigated in South Africa. The aims of this study are
115 therefore:

- 116 1. To establish the relative contributions of transpiration, canopy interception
117 and litter interception to total evaporation in an indigenous *P. henkelii*
118 even aged plantation.
- 119 2. Determine the productive and non-productive green water-use of *P.*
120 *henkelii* and *P. patula*.
- 121 3. Calculate the total green WUE of *P. henkelii* and *P. patula*.
- 122 4. Compare the total green WUE of *P. henkelii* and *P. patula*.

123

124 **METHODOLOGY**

125

126 Continuous sap flow (transpiration) monitoring on an hourly basis was
127 undertaken for both *P. henkelii* and *P. patula* for one year to incorporate
128 seasonal variations and responses to climatic factors. Event-based
129 measurements of canopy and litter interception were recorded for *P. henkelii* as
130 well. Hourly measurements of solar radiation, temperature, relative humidity,
131 wind speed and rainfall complemented these measurements and were used as
132 an input into the Variable Storage Gash canopy interception model and drying
133 curve litter interception model (Bulcock and Jewitt, 2012b), which was used to
134 estimate interception from the *P. patula* stand. The components of productive
135 and non-productive green water used to calculate total green water-use
136 efficiency is shown in **Figure 1**.

137

138 **INSERT FIGURE 1**

139 **Site Description**

140

141 Two study sites situated 50m apart from each other were selected on the Mondli
142 owned Tetworth estate in Karkloof, near Howick in the KwaZulu-Natal Midlands
143 (S 29° 21' 25.2" and E 30° 11' 49.3", alt. 1148 m.a.s.l). The sites have a mean
144 annual precipitation of 1271 mm (Lynch and Schulze, 2006), most of which falls
145 during the summer months between October and April. The mean annual
146 potential evaporation of the area is high at between 1600-1800 mm (Schulze,
147 1997), but the FAO-56 short grass reference evaporation (Allen *et al.*, 1998) was
148 1518mm for the period of the study. Mucina and Rutherford (2006) describe the
149 area as southern mistbelt forest. A characteristic of the local rainfall regime is a
150 strong orographic effect, caused by the lifting and convective cooling of the
151 summer south-east winds over the Karkloof mountain range (Dye *et al.*, 2008).
152 During the 547 day study period from October 2009 to March 2011, a total of
153 1635.9mm of rainfall was recorded, of which 346 days had rainfall. Of the 346
154 rain days, 147 (42%) of them had less than 1mm as shown in **Figure 2**. The first
155 site was a small (<1ha) even aged indigenous *Podocarpus henkelii* plantation that
156 was located close to a riparian area, but at an elevation about which the soil
157 would be classified as a true riparian area. *Podocarpus henkelii* are evergreen
158 and have long slender drooping leaves (Palgrave, 2002) which is important when
159 considering the canopy water holding characteristics and canopy interception.
160 Information on past stand management are limited due to changes in ownership
161 of this particular farm, and the trees are of an unknown age. However, by virtue

162 of their size (average tree height of 8m) and stem diameters (DBH: 15-30cm) the
163 trees are estimated to be roughly 40 years old. The trees were hand-planted, but
164 have not been pruned. The planting spacing is somewhat irregular but an
165 average distance between trees (3 m x 3 m) translates into a planting density of
166 approximately 1111 trees per hectare. The second site is a commercial
167 plantation stand of *P. patula* grown for saw timber with a planting density of 816
168 trees per hectare (3.5m x 3.5m) and is not situated in a riparian zone. They were
169 planted in September 2002, making them approximately 9 years old at the
170 beginning of the study. They have been subjected to limited thinning or pruning
171 since 2003. The soils at both sites are classified as Clovelly, characterised by an
172 orthic A-horizon and a yellow-brown apedal B-horizon with a sandy-clay-loam
173 texture and a depth of more than 1.2m. Both sites have a similar slope of 23%
174 (1:4.1) and 26% (1:3.8) for the *P. patula* and *P. henkelii* stands respectively. The
175 *P. patula* stand is on a south-west (SW) facing slope and the *P. henkelii* is on a
176 south-south-west (SSW) facing slope.

177

178 **INSERT FIGURE 2**

179 **Canopy Interception**

180

181 Throughfall measurements were undertaken at the *P. henkelii* site using a nest of
182 three “V” shaped troughs based on the design of Cuartus *et al.*, (2007)
183 constructed from galvanised sheeting (Bulcock and Jewitt, 2012a). The
184 dimensions of each trough are 0.1 m wide x 2.0 m long. Conventional “U” or “V”

185 shaped troughs were susceptible to blockage by fallen debris and water loss
186 from splash. However, this system minimizes splash out by using steep “V”
187 shaped sides. The troughs were covered with mosquito netting to minimize the
188 entry of debris, which reduced the demand of cleaning and maintaining the
189 system. A correction factor for each trough was derived from laboratory
190 measurements to account for the “initial abstraction” from the netting. The
191 troughs were then connected to a tipping bucket gauge and an event data logger.
192 Because the trough represents a linear and continuous sampling surface, the
193 linear variation of leaves, branches, and tree crown is assumed to provide a
194 representative integral of the throughfall caught (Cuartus *et al.*, (2007). The three
195 troughs were arranged in a radial pattern with an equal spacing of 120°, and
196 extended from the tree trunk towards the edged of the canopy.

197 **Litter Interception**

198

199 The litter interception and water that drains to the soil were measured at the *P.*
200 *henkelii* site using two round galvanized iron basins that fit into each other. The
201 upper basin which has a diameter of 500mm is filled with litter and has a
202 geotextile lining on top of a wire mesh base, so water can percolate into the
203 lower basin. The water that is collected in the lower basin drains into a tipping
204 bucket and records the water that would have drained to the soil. The litter
205 interception is then calculated as the difference between throughfall and the
206 water that drained to the soil. Further details of the design can be found in
207 Bulcock and Jewitt (2012a).

208

209 **Canopy and Litter Interception Models**

210

211 Due to the unavailability of canopy and litter interception data for the *P. patula*
212 stand at the Karkloof site, the “variable storage Gash model” and idealised drying
213 curve models (Bulcock and Jewitt, 2012b) were used to simulate canopy and
214 litter interception respectively for *P. patula*. The model descriptions and
215 verification are detailed in Bulcock and Jewitt (2012b).

216

217 **Sap Flow Measurements**

218

219 The Heat Pulse Velocity (HPV) technique is an internationally accepted method
220 for measuring the flow of sap in trees and has received much attention by
221 researchers in recent years, (Smith and Allan, 1996; Gush and Dye, 2009). The
222 HPV technique was used to measure the sapflow/transpiration for both *P.*
223 *henkelii* and *P. patula*. The HPV technique has been extensively applied in South
224 Africa (Dye & Olbrich, 1993; Dye, 1996; Dye, Soko & Poulter, 1996; Dye *et al.*,
225 1996; Gush, 2008; Gush & Dye, 2009) on both indigenous tree species and
226 commercial forestry species. The HPV measurements described in this paper are
227 based on the heat ratio method (HRM) described by Burgess *et al.* (2001)
228 because of its ability to accurately measure low rates of sap flow that were
229 expected in the indigenous *P. henkelii* stand. The HRM requires a line-heater to

230 be inserted in the xylem at the vertical midpoint (commonly 5 mm) between two
231 temperature sensors (thermocouples). Heat pulses are used as a tracer, which is
232 carried by the flow of sap up the stem. This allows the velocity of individual heat
233 pulses to be determined by recording the ratio of the increase in temperature
234 measured by the thermocouples (TC's), following the release of a pulse of heat
235 by the line heater. For these measurements TC pairs and heater probes were
236 positioned 80cm up the main stem of each tree, below the first branches. TC's
237 were inserted to four different depths within the sapwood to determine radial
238 variations in sap flow (i.e. each tree contained four TC pairs). The insertion
239 depths of the TC's were calculated after first determining the total sapwood depth
240 for each species, and then spacing the probes evenly throughout. While
241 performing the drilling, a drill guide was strapped to the tree, to ensure that the
242 holes were as close to parallel as possible. CR1000 data loggers connected to
243 AM16/32 multiplexers (Campbell Scientific, Logan, UT) were programmed to
244 initiate the heat pulses and record hourly data from the respective TC pairs.

245

246 Heat pulse velocities derived using the HRM were corrected for sapwood
247 wounding caused by the drilling procedure, using wound correction coefficients
248 described by Swanson & Whitfield (1981) based on the wound widths reported in
249 **Table 1**. The corrected heat pulse velocities were then converted to sap flux
250 densities according to the method described by Marshall (1958). Finally, the sap
251 flux densities were converted to whole-tree total sap flow by calculating the sum
252 of the products of sap flux density and cross-sectional area for individual tree

253 stem annuli (determined by below-bark individual probe insertion depths and
254 sapwood depth). Hourly sap flow values were recorded from all the trees.
255 Periods of missing data were patched using data from another probe set which is
256 the most highly correlated. The complete record was aggregated into daily,
257 monthly and annual totals. Individual tree sap flow volumes ($L \cdot month^{-1}$) were
258 scaled up to a hectare using the planting density to also derive sap flow
259 (transpiration) totals in mm-equivalent volumes (Gush *et al.*, 2011).

260

261 The sap flow and stem diameter increment rates were measured by Gush *et al.*,
262 (2011) in two trees at each site (**Table 1**). The study trees were selected after
263 doing an initial survey of the range of stem diameters in the plantations. Based
264 on the findings of the initial survey, trees that fell within the two most prominent
265 stem diameter classes were selected in order to be as representative of the
266 stand as possible with the limited number of samples used. Bark thicknesses of
267 the sample trees were determined by excising bark sections from the stems.
268 Measurements of sapwood depth, required to determine the insertion depths of
269 thermocouple probes for water-use measurements, were obtained using a 5mm
270 inside-diameter increment corer (Haglöf, Sweden). Cores were subsequently
271 analysed for sapwood depth using measurements of the visual distinction
272 between lighter coloured sapwood and darker coloured heartwood. Wood density
273 for the two tree species was determined using mass and volume measurements
274 (Archimedes Principle) on stem-wood samples chiselled from the trees at a
275 height near to where the probes were inserted. Monitoring began on 13 August

276 2009 and continued until the end of March 2011. The canopy and litter
277 interception monitoring began at the beginning of the wet season in October
278 2009.

279 **INSERT TABLE 1**

280

281 **Stem Growth Measurements**

282

283 In addition to sap flow measurements, stem biomass increments surveys were
284 undertaken for both *P. henkelii* and *P. patula*, in order to calculate WUE. Stem
285 volume increment measurements were carried out at the inception of the study
286 on the 13 August 2009 and subsequently a year later on the 12 August 2010 in
287 order to include all seasons in 1-year. Stem circumferences were measured at
288 1m intervals up the tree, and subsequently converted into volume by assuming
289 that the stem consists of a series of truncated cones with a complete cone on the
290 top (Gush *et al.*, 2011). The volumes (V) (m^3) of the individual cones was
291 calculated using Equation 1.

292

$$293 \quad V = (\pi \cdot r^2 \cdot h) / 3 \quad [1]$$

294

295 Where, r is the radius at the base of the cone (m), and h is the height of each
296 cone (m). The volumes of the truncated cones were calculated using Equation 2.

297

$$298 \quad V = [\pi \cdot h (r_1^2 + r_1 r_2 + r_2^2)] / 3 \quad [2]$$

299

300 Where, r_1 is the radius at the base of the truncated cone (m), r_2 is the top of the
301 truncated cone (m), and h is the height of the truncated cone (m).

302

303 The stem volume increments were converted to dry mass using the wood
304 densities determined from samples collected for each species in this study as
305 shown in **Table 1**.

306

307 **Plant Area Index Measurements**

308

309 The single-sided plant area index (PAI) (including leaves, branches and twigs)
310 was measured using the LI-COR LAI-2000 plant canopy analyser (LAI-2000, LI-
311 COR, Inc., Lincoln, Nebraska, USA). Ten sets of four readings were taken for
312 each tree. A sunlit canopy was avoided by taking the readings just before sunset
313 when the solar elevation is low (below 45°). A 45° view restrictor was used to
314 block the sensor in the field of view of the operator. This procedure was followed
315 for all sites, and the values are shown in **Table 1**.

316

317

318

319

320

321

322 **RESULTS**

323

324 The results discussed in the subsequent sections are for the period of October
325 2009 to March 2011 at the *P. henkelii* and *P. patula* stands in Karkloof.

326

327 **Relative Contributions of Transpiration, Canopy and Litter Interception.**

328

329 The transpiration recorded in the *P. henkelii* stand shows a relatively consistent
330 rate throughout the year (**Figure 3**) with monthly transpiration varying between
331 26.8mm and 48.8mm. This may be attributed to the evergreen nature of *P.*
332 *henkelii*, as well as the lack of seasonal water stress due to the riparian location
333 of the site. Also, *Podocarpus* is a gymnosperm and therefore the xylem consists
334 of tracheids which have a lower conductivity than the vessels of angiosperms.
335 The highest sap flows were recorded during the summer months when leaf area,
336 temperature and available water are high, as well as having longer day lengths.
337 Transpiration accounts for the largest water-use at 41% and 95.9% of the gross
338 precipitation for *P. henkelii* and *P. patula* respectively during the study period
339 (**Table 2 and Table 3**). It is also important to note that the transpiration by the *P.*
340 *patula* exceeds the total quantity of water that infiltrates into the soil. Therefore,
341 the *P. patula* is reliant on additional water from upslope to recharge the soil
342 water, or else is accessing ground water. *P. henkelii* transpires on average 97.8
343 L per unit of plant area per year, compared to the 436.9 L per unit of plant area
344 for *P. patula*. Therefore, even though *P. henkelii* has a smaller PAI than *P.*

345 *patula*, the exotic *P. patula* transpires on average 4.5 times more per unit of plant
346 area than indigenous *P. henkelii*. Canopy interception is the second highest
347 water-use at 29.8% and 22.1% of gross precipitation for *P. henkelii* and *P. patula*
348 respectively. The highest absolute monthly canopy interception loss for both *P.*
349 *henkelii* and *P. patula* was recorded in December 2009 at 50.4mm and 37.2mm
350 respectively. The highest canopy interception losses are expected during the
351 summer months when there is the highest rainfall, as well as highest evaporation
352 potential due to the higher temperatures. Conversely, the lowest absolute canopy
353 interception losses are recorded during the winter months when there is very little
354 rainfall, with as little as 3.7mm and 2.7mm being lost to canopy interception in
355 May 2010 for *P. henkelii* and *P. patula* respectively. Litter interception is the
356 lowest evaporative loss, accounting for only 6.2% and 10.7% of gross
357 precipitation for *P. henkelii* and *P. patula* respectively. The small litter interception
358 amount can be attributed to the large number of consecutive rain days, during
359 the rainy summer months, thereby not allowing time for much evaporation to take
360 place. The study site is also situated in a mistbelt, and the presence of mist
361 suppresses evaporation. The trees also have a dense canopy with a PAI of
362 between 3.5 and 4.0 for *P. henkelii* and between 2.3 and 2.5 for *P. patula*, and
363 therefore little solar radiation reaches the litter to aid in evaporation. During the
364 winter months, there is little rainfall, and after canopy interception losses have
365 been accounted for, there is little throughfall to be intercepted by the litter.

366 **INSERT FIGURE 3.**

367 **INSERT TABLE 2.**

368 **INSERT TABLE 3.**

369 **Productive Green Water-Use Efficiency**

370

371 The stem growth and WUE for the two *P. henkelii* and *P. patula* trees was
372 calculated for the one year period 13 August 2009 to 12 August 2010 (Gush *et*
373 *al.*, 2011) and is summarised in **Table 4** and **Table 5**. The WUE was calculated
374 as the increase in stem wood dry mass relative to transpiration. The WUE was
375 also calculated as a mm-equivalent by considering the planting density of 1111
376 and 816 stems per hectare for *P. henkelii* and *P. patula* respectively. From **Table**
377 **4** and **Table 5** it can be seen that the average productive WUE of the two *P.*
378 *henkelii* trees is 0.79 g.L⁻¹ or 7.14g.mm⁻¹ transpired and 2.06g.L⁻¹ or 25.21g.mm⁻¹
379 for *P. patula*. Dye *et al.*, (2008) found comparable WUE values of 2.40 and
380 2.50g.L⁻¹ for an 8 and 16 year old *P. patula* stand in KwaZulu-Natal respectively.
381 However, WUE values ranged from 1.40 to 4.50g.L⁻¹ depending on age and site
382 location. No other WUE studies on *P. henkelii* could be found, but a number of
383 studies on *Podocarpus falcatus* have been done. Dye *et al.*, (2008) found the
384 WUE of *P. falcatus* under plantation conditions in Magoebaskloof, Limpopo to be
385 0.86g.L⁻¹ and 1.05 g.L⁻¹ in a single tree site in Karkloof, KwaZulu-Natal. In Table
386 4 and Table 5 the values of annual transpiration are given as 195.0 mm and
387 559.2 mm for the two *P. henkelii* trees and 803.7 mm and 1311.1 mm for the two
388 *P. patula* trees. This gives a mean for the two species of 378.6 mm and 1057.4
389 mm for *P. henkelii* and *P. patula* respectively. Despite the difference in the mean
390 transpiration for *P. patula* and *P. henkelii* being large, by performing a one-tailed

391 t-test it is shown to be not statistically significant at the 95% level ($p=0.08$).
392 However, the difference in WUE between the two species is significant at the
393 95% level ($p = 0.036$), assuming that these differences are based on species and
394 not site factors.

395

396 **INSERT TABLE 4**

397 **INSERT TABLE 5**

398

399 **Total Green Water-Use Efficiency**

400

401 Typically, WUE will be reported as it is calculated above. However, it is important
402 to also consider WUE for the total water-use. Using the average mm-equivalent
403 productive WUE of the two *P. henkelii* and *P. patula* trees of $7.14\text{g}\cdot\text{mm}^{-1}$ and
404 $25.21\text{g}\cdot\text{mm}^{-1}$ transpired water respectively, as shown in **Table 4** and **Table 5**, the
405 stem mass increment for the period October 2009 to March 2011 can be
406 estimated and therefore, the total WUE was determined as shown in **Table 6**.

407 **INSERT TABLE 6**

408

409 After calculating the average stem mass increment for the period of October
410 2009 to March 2011, the total WUE was calculated by multiplying the average
411 productive WUE by the transpiration. In order to calculate the total WUE, the
412 stem mass increment was divided by the sum of all water fluxes (transpiration,
413 canopy and litter interception). When the total WUE is calculated by considering

414 all the water fluxes, the WUE of *P. henkelii* is $3.8\text{g}\cdot\text{mm}^{-1}$ as opposed to
415 $7.14\text{g}\cdot\text{mm}^{-1}$, which is a difference of 46.8%. Similarly, the total WUE of *P. patula*
416 is $18.8\text{g}\cdot\text{mm}^{-1}$ as opposed to $25.21\text{g}\cdot\text{mm}^{-1}$, which is a difference of 26.2%.

417

418 **Uncertainties and Potential Sources of Error**

419

420 One of the limitations of this study is the relatively short study period and it is well
421 documented that annual growth increments are variable from year-to-year. The
422 growth of the tree is largely seasonal, so an annual increment which includes
423 both a summer and winter period may not be accurate in upscaling for just
424 another summer period. The sample size of this study is small and was limited by
425 the availability and cost of equipment. Therefore, the variability of water-use
426 between the trees in the stand could not be well represented.

427

428 The canopy interception was not measured in the *P. patula* stand but was
429 modelled using the Variable Storage Gash model (Bulcock and Jewitt, 2012).
430 However, the Variable Storage Gash model was developed and validated in a *P.*
431 *patula* stand in the KwaZulu-Natal Midland and was found to estimate canopy
432 interception in *P. patula* well.

433

434 Stemflow was not measured, so the throughfall values are slightly high and
435 should be taken into consideration when using these results. In a review by
436 Levia and Frost (2003) of stemflow studies of different species and in different

437 locations, it was found that on average stemflow accounted for 5.5% of the gross
438 precipitation. This is similar to the findings of José (2013) whose review of
439 stemflow studies suggests that on average stemflow accounts for less than 5% of
440 gross precipitation. Therefore, if a value of 5% of gross precipitation is used as
441 an estimate of stemflow for *P. patula* and *P. henkelii*, then the throughfall will
442 effectively increase by 81.8mm for the period of this study (i.e. canopy
443 interception decreased by 81.8mm). If this value is used to recalculate the total
444 green WUE, then the total green WUE will increase to $4.1\text{g}\cdot\text{mm}^{-1}$ and $19.6\text{g}\cdot\text{mm}^{-1}$
445 for *P. henkelii* and *P. patula* respectively.

446

447 The HPV technique is the most common method of measuring
448 sapflow/transpiration and has shown to measure accurately in a variety of
449 hardwood trees. However, Pine trees have a strongly defined ring structure in the
450 sapwood which gives rise to a complex radial pattern of sapflow. In a study by
451 Dye *et al.*, (1996) which evaluated the HPV technique in *P. patula*, it was found
452 that the sapflow was overestimated by as much 49% compared to cut tree
453 uptake measurements. However, the sapflow results of this study corresponded
454 well with other studies mentioned previously. It is therefore important to bear
455 these considerations in mind when using the results of this study and it is
456 recommended that a follow-up study is undertaken where the total evaporation is
457 validated using long-term measurement with a scintillometry system or other
458 above-canopy measurement system.

459

460 **DISCUSSION AND CONCLUSION**

461

462 Many WUE studies express water-use in terms of an increase in stem wood
463 biomass relative to transpiration (productive Green water-use). While this
464 approach is useful in terms of a physiological water-use, it may be misleading for
465 water resources management and planning, particularly in areas where
466 interception by the canopy and litter are significant as has been shown to be the
467 case in the KwaZulu-Natal Midlands (Bulcock and Jewitt, 2012a). For example,
468 two different crops/trees may have similar productive Green WUE's, but one may
469 have a significantly higher or lower canopy and litter interception than the other
470 resulting in a different total Green WUE. As shown in the results of this research,
471 the difference between productive Green WUE and total green WUE where the
472 non-productive Green water fluxes are included is 46.8% and 26.2% for *P.*
473 *henkellii* and *P. patula* respectively. Therefore, for the total Green WUE approach
474 to be implemented in more studies, there is a need for sound canopy and litter
475 interception models that make use of readily available data that can be used in
476 cases where interception data are not available.

477

478 In terms of productive Green WUE, introduced species such as *P. patula* may be
479 2-4 times more efficient in their water-use than *P. henkellii*, based on the results
480 of other studies (Olbrich *et al.*, 1996; Dye *et al.*, 2001; Gush and Dye, 2009;
481 Gush *et al.*, 2011), which correspond well with the findings of this study. These
482 previous studies on the WUE of introduced species such as *P. patula* and *E.*

483 *grandis* in South Africa show that their WUE can be as high as 4.5g.L⁻¹ and
484 5.5g.L⁻¹ respectively. Gush *et al.*, (2011) reported a mean WUE from a number of
485 studies for *P. patula* to be 2.5g.L⁻¹, which is consistent with the findings of this
486 study. Gush and Dye (2009) found that the WUE of a number of indigenous tree
487 species including *Trema orientalis*, *Celtis Africana*, *Podocarpus falcatus*,
488 *Ptaeroxylon obliquum*, *Olea eurpaea* subsp. *africana* and *Berchemia zeyheri* to
489 be 0.96, 1.57, 1.04, 1.32, 0.31 and 1.67g.L⁻¹ respectively, highlighting that
490 introduced species are generally more water-use efficient than indigenous
491 species. The difference in productive WUE was found to be statistically
492 significant at the 95% level ($p = 0.036$) if it is assumed that the differences are
493 based on species and not site factors. The site factors are however very similar
494 and this is probably a fair assumption. Although the *P. henkelii* site is closer to
495 the riparian area than the *P. patula*, the soil properties at the depth at which the
496 soil profile was classified (1.2m) were the same. While the indigenous *P.*
497 *henkelii* may not be as water-use efficient as some introduced plantation species,
498 it does have a relatively lower water-use year on year. It was found in this study
499 that *P. patula* transpires on average 4.5 times more per unit of plant area than *P.*
500 *henkelii*. This large difference may be attributed to a number of physiological
501 factors. Firstly, the stomatal conductance in *Pinus* is larger than in *Podocarpus*.
502 Rolando (2008) measured maximum stomatal conductance in *P. patula* to be 150
503 mmol.m⁻².s⁻¹ and Saugier *et al.*, (1997) measured a maximum stomatal
504 conductance of 124 mmol.m⁻².s⁻¹ in *Pinus banksiana*. In comparison, Dye *et al.*,
505 (2008) measured a maximum stomatal conductance of 32.1 mmol.m⁻².s⁻¹ in

506 *Podocarpus falcatus*. Secondly, Baldocchi *et al.*, (1987) found that canopy
507 transpiration and canopy photosynthesis were strongly coupled. Myers *et al.*,
508 (1999) reported net photosynthetic rates of between 5.24 and 6.47 mmol.m⁻².s⁻¹
509 for *Pinus taeda*. In comparison, rates of between 2.34 and 3.88 mmol.m⁻².s⁻¹ for
510 *P. falcatus* were reported by Dye *et al.*, (2008) depending on photosynthetically
511 active radiation (PAR). As a result of the much higher transpiration rates by *P.*
512 *patula*, the water-use by the *P. patula* exceeds the amount of rainfall reaching the
513 soil. Conversely, the *P. henkelii* uses less water than that which drains to the soil
514 and therefore has less impact on the water resources. This is important from a
515 water resources management perspective, where the harvesting of timber is of
516 secondary importance, as the indigenous species may have a lower annual
517 reduction in streamflow than introduced plantation species. It must however be
518 stressed that *P. henkelii* cannot be used to characterise the water-use for all
519 indigenous species, as not all indigenous species will necessarily have a lower
520 water-use than commercial forestry species. *P. henkelii* is a gymnosperm with a
521 tracheid dominated xylem, and so the sapflow is slow. From a hydrological or
522 water management point of view, a potential application of low water-use
523 indigenous species such as *P. henkelii* could be to plant them in riparian areas
524 within commercially afforested areas as is the case at the site of this study. Many
525 of the narrow riparian areas of grassland that remain after commercial
526 afforestation are heavily infested with alien invasive species due to the difficulty
527 in managing them. As it is dangerous to perform bi-annual burns within the
528 plantation, it may be a viable land-use option to plant indigenous trees species in

529 these areas due to their low water-use (Gush *et al.*, 2011). However, when
530 considering the impact of planting trees within a riparian area, the productive and
531 non-productive Green water-use should be considered, as it is the total Green
532 water-use that will ultimately determine the streamflow reduction as well as the
533 water resource management and planning decision.

534

535 While it is the total Green water that needs to be considered from a water
536 resources management and planning perspective, one cannot lose sight of the
537 importance of considering the individual components of productive and non-
538 productive green water. The non-productive components of total evaporation
539 have been referred to by some as “white water” (Savenije, 2004), highlighting
540 that hydrologically it is problematic to lump these two components together and
541 that there needs to be clear recognition that these components need to be
542 considered separately in hydrological process studies. Failure to have a sound
543 conceptual understanding of the individual components that make up total Green
544 water flows may lead to modelling efforts being compromised (Jewitt, 2006). This
545 point is emphasised by Savenije (2004) who states that the “common mistake of
546 lumping interception with transpiration leads to an over-dimensioning of the soil
547 moisture stock”. Therefore, this paper highlights the importance of considering
548 the individual components of both productive and non-productive Green water
549 flows, and in particular, the role that interception plays in the hydrological cycle
550 and that from a water resources management and planning point of view, the
551 total Green WUE needs to be considered. However, it is still vitally important to

552 understand the productive water-use efficiency for the optimisation of future
553 water, food and timber requirements.

554 **Acknowledgements**

555

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561 CSIR is appreciated. Mr Mark Norris-Rogers of Mondi is thanked for his
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730 **FIGURES**

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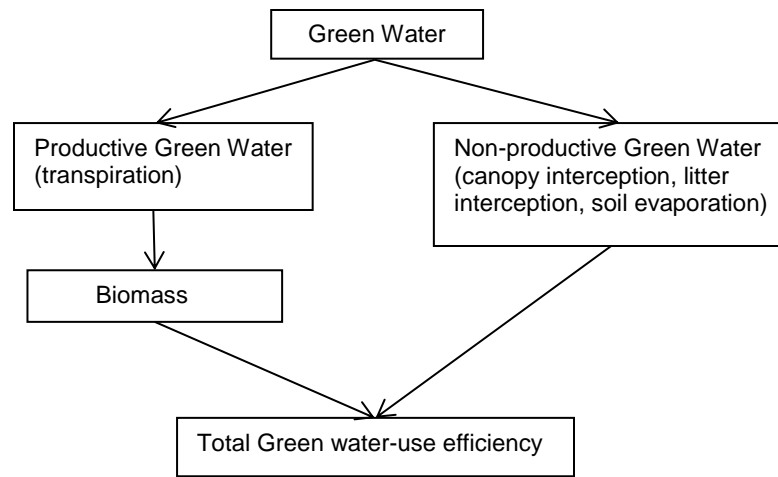
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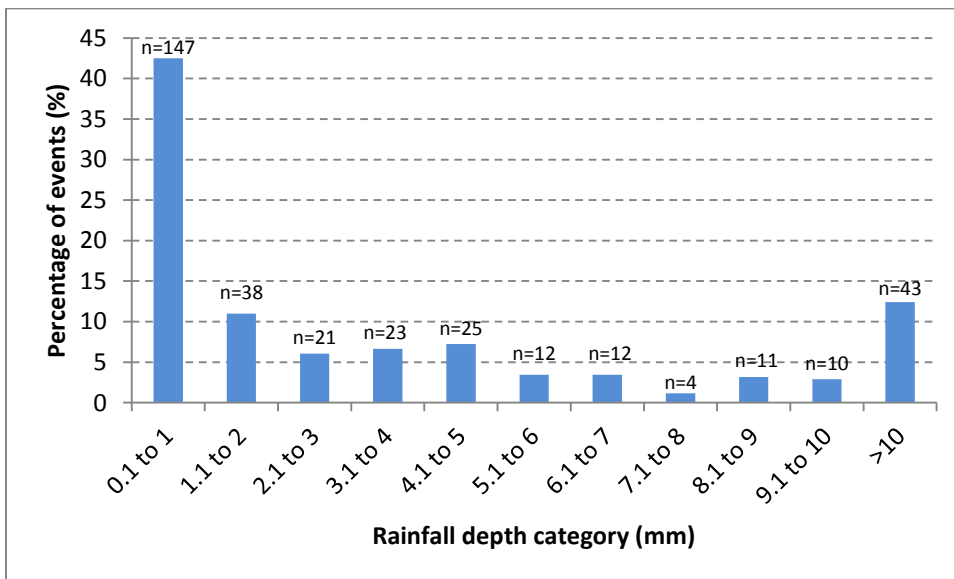
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740 Figure 1. Schematic showing the components of productive and non-productive
741 green water used to calculate total green water-use efficiency.

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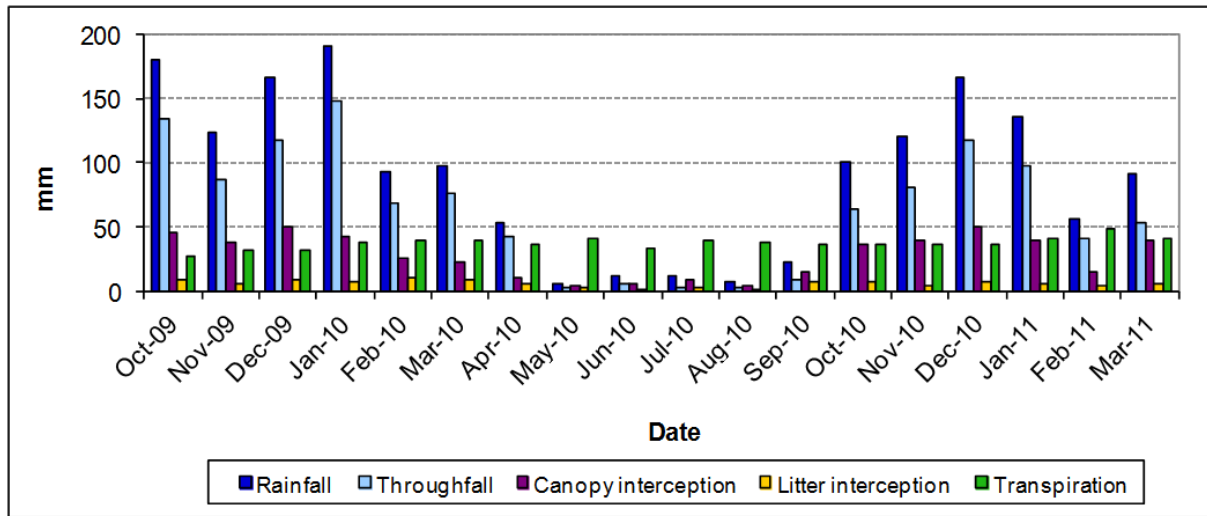
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744 Figure 2. Percentage of rainfall events per rainfall depth category for the period

745 October 2009 to March 2011

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749 Figure 3. Measured contributions of rainfall, throughfall, canopy interception, litter
750 interception and transpiration for the period October 2009 to March 2011 for *P.*
751 *henkelii*.

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765 **LIST OF TABLES**

766

767 Table 1. Sample tree details as of 12 August 2010.

Tree	Diameter at Breast height (mm)	Tree height (m)	PAI	Sapwood depth (mm)	Wound width (mm)	Bark width (mm)	Wood density (g.cm ⁻³)	Mean litter thickness (mm)
<i>P. henkelii</i> 1	140	6.34	3.5	55	3	7	0.468	52
<i>P. henkelii</i> 2	230	7.33	4.0	95	3	7	0.468	52
<i>P. patula</i> 1	200	8.77	2.3	85	4	10	0.380	151
<i>P. patula</i> 2	240	10.79	2.5	100	4	10	0.380	151

768

769 Table 2. Monthly contributions and totals of rainfall, throughfall, canopy

770 interception, litter interception and transpiration measured for the period October

771 2009 to March 2011 for *P. henkelii*.

Date	Rainfall (mm)	Throughfall (mm)	Observed canopy interception (mm)	Observed litter interception (mm)	Water drained to soil (mm)	Transpiration (mm) <i>P. henkelii</i>
Oct-09	179.6	134.5	45.1	8.9	125.6	26.8
Nov-09	123.6	86.2	37.4	5.2	81.0	32.2
Dec-09	167.2	116.8	50.4	9.0	107.8	32.3
Jan-10	190.6	148.8	41.8	7.3	141.5	37.2
Feb-10	93.1	68.0	25.1	9.9	58.1	39.0
Mar-10	98.3	76.3	22.0	9.2	67.1	39.1
Apr-10	53	42.5	10.5	6.2	36.3	36.2
May-10	6.1	2.4	3.7	1.9	0.5	40.3
Jun-10	11.6	5.7	5.9	1.5	4.2	33.8
Jul-10	11.5	3.3	8.2	2.0	1.3	39.0
Aug-10	7.6	2.7	4.9	1.0	1.7	38.3
Sep-10	22.1	8.2	14.9	7.0	1.2	36.2
Oct-10	100.4	63.7	36.7	7.9	55.8	36.2
Nov-10	120.2	80.7	39.5	3.6	77.1	36.7
Dec-10	166.4	116.9	49.5	7.3	109.6	36.0
Jan-11	135.9	97.1	38.8	5.3	91.8	41.3
Feb-11	56.6	41.5	15.1	3.5	38.0	48.8
Mar-11	92.1	53.4	38.7	5.1	48.3	40.8
Total (mm)	1635.9	1148.8	488.1	101.8	1047.0	670.3
Percentage of rainfall (%)		70.2	29.8	6.2	64.0	41.0

772

773 Table3. Monthly contributions and totals of rainfall, throughfall, modelled canopy
 774 interception, modelled litter interception and transpiration for the period October
 775 2009 to March 2011 for *P. patula*.

776

Date	Rainfall (mm)	Throughfall (mm)	Modelled canopy interception (mm)	Modelled litter interception (mm)	Water drained to soil (mm)	Transpiration (mm) <i>P.</i> <i>patula</i>
Oct-09	179.6	146.3	33.3	17.3	129	66.0
Nov-09	123.6	96.0	27.6	10.2	85.8	73.5
Dec-09	167.2	130.0	37.2	17.4	112.6	82.7
Jan-10	190.6	159.7	30.9	14.2	145.5	94.3
Feb-10	93.1	74.6	18.5	9.2	65.4	124.2
Mar-10	98.3	82.1	16.2	10.8	71.3	113.6
Apr-10	53	45.2	7.8	12.0	33.2	93.9
May-10	6.1	3.4	2.7	1.7	1.7	104.2
Jun-10	11.6	7.2	4.4	2.9	4.3	67.3
Jul-10	11.5	5.4	6.1	3.9	1.5	62.5
Aug-10	7.6	4.0	3.6	1.9	2.1	79.6
Sep-10	22.1	11.1	11.0	9.5	1.6	74.8
Oct-10	100.4	73.3	27.1	15.3	58	71.7
Nov-10	120.2	90.8	29.2	7.0	83.8	79.6
Dec-10	166.4	129.9	36.5	14.1	115.8	76.7
Jan-11	135.9	107.3	28.6	10.3	97	94.7
Feb-11	56.6	45.5	11.1	6.9	38.6	109.4
Mar-11	92.1	63.5	28.6	9.9	53.6	101.1
Total (mm)	1635.9	1275.3	360.4	174.5	1100.8	1569.8
Percentage of rainfall (%)		77.9	22.1	10.7	67.3	95.9

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783 Table 4. Summary of productive WUE data for *P. henkelii* trees as calculated
 784 from a mass-based ratio of biomass increment relative to productive green water-
 785 use for the one year period 13 August 2009 to 12 August 2010.

Tree	1yr water-use (L)	1yr water-use (mm)	Stem Volume increment (m ³)	Wood Density (g.cm ⁻³)	Stem mass increment (g)	WUE (g stem wood.L transpired water ⁻¹)	WUE (g stem wood.mm transpired water ⁻¹)
<i>P. henkelii</i> 1	1755	195.0	0.00215	0.468	1006.2	0.5733	5.16
<i>P. henkelii</i> 2	5033	559.2	0.01088	0.468	5091.8	1.0117	9.11
Average	3394	378.6	0.00652	0.468	3049.0	0.7925	7.14

786

787

788 Table 5. Summary of productive WUE data for *P. patula* trees as calculated from
 789 a mass-based ratio of biomass increment relative to productive green water-use
 790 for the one year period 13 August 2009 to 12 August 2010.

Tree	1yr water-use (L)	1yr water-use (mm)	Stem Volume increment (m ³)	Wood Density (g.cm ⁻³)	Stem mass increment (g)	WUE (g stem wood.L transpired water ⁻¹)	WUE (g stem wood.mm transpired water ⁻¹)
<i>P. patula</i> 1	9849	803.7	0.05157	0.380	19596.6	1.9897	24.23
<i>P. patula</i> 2	16067	1311.1	0.09035	0.380	34333.0	2.1369	26.19
Average	12958	1057.4	0.07096	0.380	26964.8	2.0633	25.21

791

792

793 Table 6. Summary of WUE data for *P. henkelii* and *P. patula* trees as calculated
 794 from a mass-based ratio of biomass increment relative to total green water-use
 795 for the period October 2009 to March 2011.

Tree	Average productive WUE (g.mm transpired ⁻¹)	Transpiration (mm)	Stem mass increment (g)	Canopy interception (mm)	Litter interception (mm)	Total Green water (mm)	Total Green WUE (g.mm total green water ⁻¹)
<i>P. henkelii</i>	7.14	670.3	4785.9	488.1	101.8	1260.2	3.8
<i>P. patula</i>	25.21	1569.8	39574.7	360.4	174.5	2104.7	18.8

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