

1 **Incorporating stakeholder preferences in the selection of technologies for using invasive**
2 **alien plants as a bio-energy feedstock: applying the analytical hierarchy process**

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25

26 **Abstract**

27

28 Invasive alien plants (IAPs) impose significant social costs on the population of the Agulhas
29 Plain region in South Africa due to their adverse impacts on ecosystem goods and services
30 (decreased water supply and increased fire risk) . While the cost of clearing IAPs is
31 considerable, this paper assesses opportunities to reduce some of the social and
32 environmental burdens (e.g. disruptions of ecosystems which have negative impacts on
33 livelihoods) by using IAP biomass to produce bio-energy. However, such an initiative could
34 increase financial dependency on these plants and is thus considered to be a major risk factor
35 which could create adverse incentives to illegally grow these plants. A participatory
36 decision-making process with active stakeholder participation is a key element in managing
37 such an initiative. We used a multi-stakeholder engagement process and the analytical
38 hierarchy process to define and weigh suitable criteria for the assessment of different “IAP
39 biomass to bio-energy” technology scenarios on the Agulhas Plain. Feasible scenarios were
40 constructed by means of an expert panel which were then ranked according to stakeholder
41 preference. The six criteria were: minimising impacts on natural resources; job creation;
42 certainty of benefits to local people in the study area; development of skills for life;
43 technology performance and cost efficiency. This ranking was largely determined by the
44 preference for resource efficiency in terms of minimising impacts on natural ecosystems and
45 the localisation of benefits. The smaller, modular technologies were consequently preferred
46 since these realise direct local benefits while developing local skills and capacity in their
47 manufacture, sales and maintenance. The rankings as obtained in this study are context-
48 bound, which implies that the findings only have limited application to areas with similar
49 biophysical and socio-economic characteristics. However, the method itself is fully
50 generalisable, and the same prioritisation process can be followed in any study area to ensure

51 that a participatory decision making process fulfils local energy needs and contributes to
52 sustainable development.

53

54 **Keywords**

55

56 invasive alien plants; bio-energy; analytical hierarchy process; stakeholder preference; multi-
57 criteria decision analysis

58

59 **1 Introduction**

60

61 The growth in human populations has been accompanied by unprecedented encroachment on
62 terrestrial ecosystems, and the expansion of global trade has led to the widespread
63 distribution of large numbers of species beyond their native ranges- this increasingly
64 threatens the integrity of ecosystems and the services they deliver (Perrings et al., 2010,
65 Pimentel, 2002, Pimentel et al., 2001, De Lange and Van Wilgen, 2010). In South Africa,
66 invasive alien plants (IAPs) in South Africa have increased sufficiently in terms of magnitude
67 and distribution (Marais et al., 2004, Marais and Wannenburgh, 2008, Mgidia et al., 2007,
68 Kotzé et al., 2010, Van Wilgen et al., 2008) to degrade biodiversity and ecosystem resilience
69 by changing the configuration of ecosystems and ecosystem-derived goods and services (De
70 Lange and Van Wilgen, 2010, Pimentel et al., 2005, Moran et al., 2005). This creates serious
71 economic losses and management challenges, since biodiversity has a direct value to society
72 through the provision of life-sustaining ecosystem goods and services; as well as ecological
73 resilience. IAPs have a significant negative impact on the reliability and security of these
74 ecosystem services, which hampers economic growth and eventually (often unwittingly)
75 degrades social welfare. Specifically, one of the most substantial impacts of IAPs in South

76 Africa is their impact on water availability, causing an estimated 6.7 % decrease in mean
77 annual runoff (Versfeld et al., 1998; Le Maitre et al, 2000). Given the arid climate and
78 delicate water balance in South Africa, these losses present a serious opportunity cost to the
79 economy. For example, one study has estimated that the monetary cost of these losses is close
80 to ZAR6 billion per annum (De Lange and Kleynhans, 2008).

81

82 South Africa has been heavily impacted by IAPs, with approximately 2.5 million condensed
83 hectares currently invaded (Kotze, 2010), however this figure is heavily debated. Globally,
84 the response to the threat of IAPs has varied, with several countries (including South Africa)
85 developing strategies for dealing with the problem in an integrated fashion, which includes
86 combinations of mechanical, chemical and biological control, and habitat management
87 (McNeeley et al., 2001). These control measures create a substantial fiscal burden to the
88 South African economy. The Working for Water programme was established by the
89 Department of Water Affairs in 1998 in order to address this problem (Turpie et al., 2008,
90 Buch and Dixon, 2008). The Working for Water programme is an expanded public works
91 programme in which IAPs are cleared using labour-intensive techniques. Funding for the
92 Programme originates from various sources including water user charges, government
93 funding aimed at poverty relief. The programme is seen as a success-story in controlling
94 IAPs in South Africa, while at the same time creating employment (Turpie et al., 2008,
95 Hobbs, 2004, Buch and Dixon, 2008). Financial benefits from the utilisation of IAP biomass
96 removed through the Programme has been limited to use of the wood for fire-wood, crafts
97 and furniture; but most of the biomass is left *in situ*, which often results in a fire hazard.
98 Therefore, additional options for utilising this biomass need to be explored.

99

100 We explored the opportunity of utilising IAPs biomass for the production of bio-energy.
101 With rapid advancements in renewable energy technologies that can utilise biomass feedstock
102 to produce bio-energy products, there is an urgent need to assess and compare the
103 appropriateness of and preference for these technology options in a given case. This study
104 developed five scenarios based on different bio-energy technology options and the bio-energy
105 services that they can provide. These options were: compressed logs, slow pyrolysis for
106 charcoal, gasification for electricity, combustion for electricity, and decentralised heat and
107 power. The analytical hierarchy process (AHP), a multi-criteria decision analysis (MCDA)
108 method, was used to determine relevant assessment criteria and to identify the preferred
109 scenario (Saaty, 1980). The ranking of scenarios was based on a number of criteria, which
110 were weighted by combining expert opinion with community preferences.

111

112 The study area was the Agulhas Plain region in the Southern Cape, South Africa. The region
113 has a land area of 270 000 hectares and a population of approximately 45 000. It lies at the
114 southern-most tip of Africa, between the ocean to the south and the Langeberg mountain
115 range to the north. The area is administered by the Overstrand and Cape Agulhas local
116 municipalities, which form part of the Overberg District Municipality (Naudé *et al.*, 2007).

117

118 Recent estimates of IAP species composition and density of invasions (Kotzé *et al.*, 2010)
119 enabled us to estimate a biomass yield of approximately 2.3 million tonnes per annum over a
120 20 year period for the Agulhas Plain (Table 1), with *Acacia saligna* and *Acacia cyclops*
121 dominating the species composition (De Lange and Le Maitre, 2010).

122

123 **2 MCDA techniques**

124

125 MCDA is an umbrella term to describe a collection of formal approaches used to facilitate
126 decision-making for complex problems where there are multiple conflicting criteria (Rozakis
127 et al., 2001, Belton and Stewart, 2002). The aim is to integrate measurement procedures with
128 value judgements in an attempt to make the inevitable subjectivity in complex problems more
129 explicit, while improving stakeholder buy-in and consequently the acceptability of decisions
130 (Eberhard and Joubert, 2002, Joubert *et al.*, 2003, Cherni *et al.*, 2007). The approach
131 facilitates a greater understanding of complex management problems by taking explicit
132 account of criteria used to explore and analyse choices made by stakeholders (priorities,
133 values and objectives) in the context of the problem; the outputs can guide decision-makers to
134 identify a preferred course of action (Mamphweli and Meyer, 2009, Strager and Rosenberger,
135 2006, Hobbs et al., 1992, Belton and Stewart, 2002). This approach provides a transparent
136 and rational methodology to guide and rank management alternatives based on weighted
137 evaluation criteria (Buchholz *et al.*, 2009, Buchholz *et al.*, 2007). It does not eliminate all
138 subjectivity in the decision-making process, but reflects the trade-offs the respondents are
139 willing to make in a specific context (Stewart *et al.*, 2001, Stewart *et al.*, 1997, Hobbs *et al.*,
140 1992). A level of subjectivity in decision-making will thus remain, but the process
141 nonetheless facilitates decision-making by making the need for subjective choice explicit, and
142 the process of taking account of this subjectivity more transparent (Stewart *et al.*, 2001,
143 Stewart *et al.*, 1997, Hobbs *et al.*, 1992, Stewart, 2004). Such transparency is important
144 because it enables stakeholder participation and builds confidence in the decision-making
145 process, especially in cases where multiple stakeholders with conflicting views are involved
146 (Starkl and Brunner, 2004).

147

148 Different approaches to MCDA can be followed (Belton and Stewart, 2002):

149 a) Goal programming and reference point techniques are regarded as the original formal form
150 of MCDA. b) Utility and value-function approaches, of which multi-attribute value function
151 (MAVF) and analytical hierarchy process (AHP) are well-known. c) Game theory
152 approaches, which aim to identify solutions that are the most acceptable compromise between
153 stakeholders.

154 The AHP method was chosen for this study since it has been widely applied to determine
155 preferences in complex, multi-attribute problems (Ananda and Herath, 2005, Hobbs et al.,
156 1992, Hwang and Yoon, 1980, Stewart et al., 1997, Belton and Stewart, 2002, Pavlikakis and
157 Tsihrintzis, 2003, Kablan, 1997, Herath, 2004, Duke and Aull-Hyde, 2002). The initial steps
158 in the AHP are to develop and obtain consensus with regard to the goal and criteria hierarchy,
159 against which suitable management alternatives are compared in pair-wise fashion (Ananda
160 and Herath, 2003b, Ananda and Herath, 2003a, Herath, 2004, Kablan, 1997). The aim is to
161 establish the relative preference order of preference for alternatives that are assessed against
162 the goal/objective. Pair-wise comparisons of alternatives simply compare two alternatives
163 with each other and record the relative preference for one above the other in terms of a given
164 criterion on a numerical or descriptive scale. Whenb comparing alternatives with respect to
165 the criterion (i.e. ‘scoring’ the alternatives), participants are requested to express their
166 preferences of alternatives only with respect to a particular criterion. If no clear preference is
167 expressed, an equal score is allocated and the participants move on to the next comparison.
168 The pair-wise comparative scores are captured in a matrix, where the strength of a preference
169 is calculated in terms of a ratio of the scores.

170
171 The weighting procedure uses the same pair-wise comparative approach to elicit the relative
172 preferences of decision-makers for the criteria. A Likert scale was used to note and aggregate
173 preferences within sub-criteria groups (sub-criteria are not compared between different

174 criteria groups). A comparison vector is then calculated to present the relative performance
175 of each criterion. The aim is to find the set of values (weights) which approximate the set of
176 ratios derived from the pair-wise comparisons. This is done by eigenvalue analysis of
177 matrices, which aims to extract the eigenvector corresponding to the maximum eigenvalue of
178 the pair-wise matrix. The procedure is iterative and software programs which implement
179 multi-criteria approaches like the AHP, such as “Expert Choice[®]” (Expert Choice, 2009) are
180 often used to facilitate the analysis. The elements of the vector of scores are then normalised,
181 to allow for the addition and deleting of criteria in a consistent fashion. The criteria weights
182 are multiplied by the score of each alternative against each criterion to present a weighted
183 score for each alternative for each criterion. The aggregated final score of all criteria for each
184 alternative is a ranking which reflects the participants’ overall preference, which can aid in
185 decision-making. The weighting procedure for criteria, and scoring procedure for the
186 alternatives, can be done in any order, as long as both are done with the set goal/objective in
187 mind. The effectiveness of the approach depends on the interaction with the participants and
188 the presentation of information in a way that facilitates active participation to enhance
189 understanding, learning and discussion.

190

191 **3 Method and data inputs for the study**

192

193 **3.1 Incorporation of stakeholders**

194

195 The Rio Declaration on Environment and Development states that: “*Environmental issues*
196 *are best handled with the participation of all concerned citizens, at the relevant level*”
197 (United Nations Department of Economic and Social Affairs: Division of Sustainable
198 Development, 2004). The underlying rationale for stakeholder participation in natural

199 resource management decision-making is to accommodate as many opinions and preferences
200 as possible, and to create stakeholder buy-in that represents a consensus in the decision-
201 making process. Public involvement in environmental decision-making is thus recognized as
202 one of the basic requirements of sustainable development, and various studies have been
203 conducted on specific methodologies aimed at effectively accommodating public preferences
204 in decision making. (Iacofano, 1990, Smith, 1984, Renn et al., 1995, Dungumaro and
205 Madulu, 2003, Doelle and Sinclair, 2006, Messner et al., 2006, Renn, 2006, Stagl, 2006,
206 Sowman et al., 2006, Charnley and Engelbert, 2005, Maguire and Lind, 2004, Schemmel,
207 2004, Buchy and Hoverman, 2000, Davis, 1996, Alvarez-Farizo and Hanley, 2002, Kline and
208 Wichelns, 1998). Such approaches include systems dynamics modelling used interactively
209 in a public forum (Stave, 2003), value function approaches (Ananda and Herath, 2003a),
210 stated preference models (Haider et al., 2002), multi-criteria approval (Laukkanen et al.,
211 2001, Dennis, 2000), direct preference investigation (Kolokytha et al., 2002), the AHP (Duke
212 and Aull-Hyde, 2002, Tzeng et al., 2002, Ananda and Herath, 2003b, Herath, 2004, Hung et
213 al., 2006, Krajnc and Glavic, 2005), photo questionnaires (Ryan, 2002), determination
214 methodologies (Figueira and Roy, 2002), and the social discount rate (Gollier, 2002, Brekke
215 and Johansson-Stenman, 2008, Kaplow, 2006, Du Preez, 2004, Arrow et al., 2003).

216

217 There are often questions and debate as to the extent that stakeholder opinion should be
218 incorporated in highly complex decision-making regarding natural resources and other public
219 goods (Wiseman et al., 2003, Maguire and Lind, 2004, Litva et al., 2002, Buchy and
220 Hoverman, 2000, Pateman, 1970, Munro-Clark, 1990). The varied literature on this issue
221 stems from two main sources: 1) political sciences, with discussions around democracy and
222 citizenship, especially within the context of regional and local planning (Pateman, 1970,
223 Munro-Clark, 1990, Davis, 1996); and 2) development theory, especially within the context

224 of sustainable land use (Wignaraja et al., 1991, Rahman, 1993, Nelson and Wright, 1995,
225 Chambers, 1997). Both of these views confirm the importance of stakeholder involvement;
226 because without stakeholder participation, decision makers assume the risk of enforcing
227 compliance on an unwilling public (Maguire and Lind, 2004). However, the appropriate
228 level of involvement is contested, because the debate regarding whether stakeholder
229 preferences have a legitimate role to play in priority setting is highly polarised - sceptics warn
230 against the dictatorship of the uninformed, while advocates proclaim the legitimacy of the
231 stakeholder participatory process (Wiseman *et al.*, 2003, Litva *et al.*, 2002).

232

233 Our approach was to incorporate broad-based public participation that facilitated public-
234 private partnerships and was supported by government policy. Therefore, the following
235 stakeholder groups were included in this study: government expanded public works
236 programmes like Working for Water and Working on Fire (see above) and Landcare
237 (Department of Agriculture - aim is sustainable management and use of agricultural natural
238 resources); the Agulhas Biodiversity Initiative (an NGO and pilot landscape initiative of the
239 Cape Action Plan for People and the Environment); land-owners and farmer-associations;
240 local community members and municipality representatives; as well as research organisations
241 (Council for Scientific and Industrial Research and University of Cape Town).

242

243 **3.2 Structuring and formulating the goal and management alternatives**

244

245 The goal for the AHP process was to establish the preferred way to enhance the eradication of
246 invasive alien plants by means of processes that utilise all and the whole of such plants (do
247 not leave the smaller trees, unlike the harvesting for other wood products such as timber) for
248 the generation of bio-energy. Furthermore, since the aim is to eradicate IAPs, it implies that

249 the biomass-to-bio-energy options will require decommissioning, relocation or a change in
250 feedstock as an exit strategy if the IAP biomass becomes depleted. Five established and
251 commercial stage technologies were identified that could satisfy this goal- compressed logs,
252 slow pyrolysis for charcoal, gasification for electricity, combustion for electricity, and
253 decentralised combined heat and power). Using a life cycle approach, five scenarios were
254 developed around these technologies, with all scenarios having the following common pre-
255 processing steps: mechanical harvesting, a short haul to stack biomass in windrows to dry in
256 the field, and chipping using a mobile drum chipper. The wood-chips are then delivered to
257 the energy plant for conversion into a bio-energy product, which may be a fuel (such as
258 compressed wood-logs or charcoal), or electricity that is fed into the electrical grid (national
259 or local mini-grid). Pre-processing (drying and chipping) creates a common platform to
260 compare the different scenarios, while increasing the ease of handling; and facilitates uniform
261 and controlled thermal conversion. Although the drying process could increase wear and
262 maintenance on chipper blades, drying of wood increases the energy density, reduces the rate
263 of wood decay, and facilitates improved thermal conversion (which typically requires <30%
264 moisture content). Specifically, drying and chipping typically results in an almost doubling
265 of the energy density; with green wood (45% moisture) having an energy density of
266 10MJ/kg, and completely dry wood-chips (0% moisture) an energy density of nearly
267 20MJ/kg (CSIR, 2010a).

268

269 Beyond the initial pre-processing, the value chains for the five scenarios proceeded as
270 follows:

271

- 272 1. *Compressed logs* - chipped biomass (ca. 30mm) is further chipped (to ca. 3-5mm),
273 dried in an oven to <10% moisture, and then used to make extruded logs. Extruded

274 logs are transported to market where they are purchased and used in the household for
275 cooking and space-heating.

276 2. *Slow pyrolysis for charcoal* - charcoal is produced in modern retort kilns and
277 processed to briquettes with an energy density of approximately 28 MJ/kg. The
278 process of slow pyrolysis to produce charcoal has an established use in domestic and
279 industrial sectors such as the smelting industry, and could replace coal with little or no
280 modification to existing coal-fired power plants. For this scenario, the chipped
281 biomass (ca. 30mm) undergoes slow pyrolysis (carbonisation), and is then ground into
282 a powder, mixed with 5-10% starch and pressed into charcoal briquettes, which are
283 dried before being distributed to the market.

284 3. *Gasification for electricity* – chipped biomass is fed directly into the gasification
285 process produces syngas which can be converted to liquid fuels, synfuels (energy
286 value of approximately 44 MJ/kg;), or coupled to a gas turbine to generate electricity
287 that can supply a mini- or national- grid. The synfuels are valuable liquid fuels that
288 can replace diesel; they have a higher cetane value and burn with fewer emissions
289 (polyaromatics and sulphur). However, this technology of biomass-to-liquids is not
290 yet at an established commercial stage. Therefore, for the purposes of this paper, this
291 scenario was based on gasification to produce syngas that is scrubbed and fed into a
292 gas turbine for electricity generation to supply the national electricity grid.

293 4. *Combustion for electricity* – chipped and dried biomass is directly combusted and
294 coupled to a steam turbine to generate electricity to supply the national electricity
295 grid.

296 5. *Combined heat and power* - chipped biomass is used locally in a small combined heat
297 and power (CHP) unit where the chips are combusted and coupled to a small turbine

298 to generate electricity (basic electricity for 2-4 lights and cell-phone, radio and
299 television); while the heat is used for cooking, hot-water and space-heating.

300

301 **3.3 Criteria identification and results of weighing and scoring procedures**

302

303 Criteria identification is a crucial step in the AHP process. Together with the stakeholders,
304 we identified six main criteria, of which two had sub-criteria. The first criterion focussed on
305 minimising the impacts of the scenario on the natural resource base. This criterion was sub-
306 divided into air emissions, solid waste, water usage and net energy balance. The decision
307 rule was that lower emissions would be preferable. Being a national priority, job creation
308 was the second main criterion, and more labour intensive scenarios were preferred. The
309 development of skills for life, and the certainty that benefits associated with the scenario will
310 reach local people, were also considered important (criteria three and four). The performance
311 of the technology employed in the scenario (criterion five) was sub-divided into the ease of
312 maintenance and reliability of the technology itself. Cost efficiency was the sixth and final
313 criterion.

314

315 Criteria weighting, carried out by means of eliciting the preference order of multiple
316 stakeholders during several workshops, was used to construct a criteria tree diagram (Figure
317 1). Expert Choice© software (Expert Choice, 2009) was used to facilitate all pair-wise
318 comparisons during the workshops and to calculate the weights of the relative preferences
319 for criteria. These weights were combined in a criteria tree diagram, which aggregated sub-
320 criteria (indicated with an “L”) to main criteria (indicated with a “G”) (see Figure 1).

321

322 The criterion related to protection of natural ecosystems was ranked highest (0.335) by
323 stakeholders, of which the energy balance and water footprint were considered to be the most
324 important sub-criteria. Job creation was considered to be the second most important
325 criterion, with a weight of 0.224, while cost efficiency was considered least important, with a
326 weight of 0.042. These weights present the relative preference of stakeholders, and were
327 used along with data solicited from various sources to present a weighted score for each
328 scenario that considers the production of the bio-energy product as well as the end-use. The
329 data for the scoring procedures is summarised in Table 2.

330

331 The above-mentioned technical data were used as data inputs in the Expert Choice[®] software
332 to score each scenario. The weights were then applied to these scores and aggregated to yield
333 the preferred ranking as presented in the sensitivity diagram. Figure 2 presents a summary of
334 the results. The weights are presented in the form of the underlying bar-chart, while the line
335 graphs represent the performance of each scenario relative to each main criterion.

336

337 Scenario 5 (decentralised combined heat and power) came out to be the preferred scenario for
338 this study area, with scenario 1 (compressed logs) the next best alternative. It was clear from
339 the study that criteria with higher weights (e.g. “minimise impacts on natural
340 resources”(33.5%), “job creation”(22.4%) and “certainty of benefits to local people”
341 (20.8%)) will have greater impacts on the final scoring as compared to, for example “cost
342 efficiency”, which had a relatively low weight (4.2%). This highlights the effect of weighting
343 the various criteria and the need for a sensitivity analysis.

344

345 **4 Sensitivities**

346

347 Sensitivity analysis is used, to explore the effect of a decision maker's uncertainty about their
348 values and priorities, or to offer a different perspective on the problem. Saaty (1980)
349 developed a consistency index which compares the scores/weights to a value derived by
350 generating random reciprocal matrices of the same size, to give a consistency ratio which is
351 meant to have the same interpretation no matter what the size of the matrix. A consistency
352 ratio of 0.1 or less is generally seen to be acceptable. It has been suggested that instead of
353 direct pair-wise comparisons of scenarios, a so-called 'absolute measurement mode' should
354 be used in order to eliminate the need to laboriously present the performance of alternatives
355 with painstaking accuracy. This 'absolute level' of performance for each criterion is
356 predefined, which facilitates the comparison of alternatives by means of relative instead of
357 absolute scores (Belton and Stewart, 2002, Saaty, 1980). An advantage of the absolute
358 measurement mode is thus that the resultant scaling of the scores for each criterion is
359 independent of the alternatives and may be similar across criteria, which enables the
360 interpretation of criteria weights more effectively. However, reported applications of AHP
361 seldom refer to the use of absolute measurement (Belton and Stewart, 2002, Hung *et al.*,
362 2006, Pavlikakis and Tsihrintzis, 2003). We carried out sensitivity in absolute measurement
363 mode by altering the weighting of the criteria, with the aim of identifying the absolute level of
364 a particular criterion that will cause a change in the ranking of scenarios as presented in
365 Figure 2. For example, "minimising impacts" will need to be decreased in relative
366 importance from 33.5% to 23.4% in order to make scenarios 5 and 1 equally preferable.
367 Likewise, "certainty of benefits" will need to be decreased in importance to 10.4% to make
368 scenarios 5 and 1 equally preferable. The relative importance of "cost efficiency" will need
369 to be increased to 21.8% in order to make scenarios 5 and 1 equally preferable. If the
370 importance of "cost efficiency" is increased to 100%, scenario 1 will become the preferred
371 scenario with scenario 5 in second place at 29%. However if "cost efficiency" is of no

372 importance (0%), the hierarchy will stay the same. “Job creation” will need to be increased
373 in importance to 29.9% to balance out scenarios 5 and 1. If the importance of “job creation”
374 is increased to 100%, scenario 1 will become the preferred scenario, scenario 2 will take
375 second place and scenario 5 will drop to last place. Lastly, “technology performance” will
376 need to be increased to 15.4% to balance scenarios 5 and 1. It should be noted that each of
377 the proposed changes will affect the whole weight structure of the criteria tree and that none
378 of these changes can be executed without a well structured argument and participant
379 agreement to the proposed change.

380

381 **5 Discussion and conclusions**

382

383 It is argued that the financial burden of clearing invasive alien plants in the Agulhas Plain
384 could be reduced by utilising IAP biomass for bio-energy. However, this could be done in
385 several ways, and choosing the most acceptable option is a complex problem which includes
386 numerous trade-offs between different stakeholders (landowners, business, government, civil
387 society and communities in the Agulhas Plain). We applied a multi-criteria decision analysis
388 decision-support process in order to assess different IAP technology options and stakeholder
389 preferences associated with each, simultaneously. A multi-stakeholder engagement process
390 (the analytical hierarchy process) was used to develop and weigh criteria that can be used to
391 guide the assessment and management of the options. The AHP is a structured approach to
392 trade-off analysis that requires active participation from stakeholders in order to facilitate
393 discussion and understanding, which helps to ensure accountability and transparency. Such
394 transparency increases stakeholder buy-in and acceptance of the outcome, simply because of
395 the nature of the participatory process. However, care must be taken to make the subjectivity
396 associated with the process as explicit as possible. AHP does this by presenting pair-wise

397 comparisons to stakeholders as a stimulus for debate, which is arguably an efficient way to
398 engage subjectivity.

399

400 Several criteria were identified to assess the different IAP to bio-energy scenarios: minimise
401 impacts on natural ecosystems, job creation, skills development, certainty of benefits to locals
402 in the Agulhas Plain, technology performance and cost efficiency. The scenarios,
403 incorporating different IAP biomass to bio-energy technology options, were assessed by
404 means of a weighted score, which revealed the following preferred order: decentralised
405 combined heat and power (CHP), compressed logs, pyrolysis for charcoal, gasification for
406 electricity, and combustion for electricity. The order was largely determined by the
407 preference for localised benefits and minimum impacts on the environment. Options scored
408 higher when benefits tended to be localised, and when impacts to the environment were
409 minimised. Smaller, modular technology approaches which are more appropriate in the
410 localised context, namely using wood-chips or compressed logs in efficient combined heat
411 and power stoves, also have the advantage of developing local skills and capacity in their
412 manufacture, sales and maintenance. Since this assessment approach considered the entire
413 value chain, changes in end-use practices can greatly influence the overall ranking of
414 scenarios. Furthermore, in practice, in any scenario in which IAPs will be utilised for a
415 specific purpose, it is important to be aware of the risk of creating adverse dependencies
416 which could lead to incentives for ‘farming’ these plants, thus perpetuating their invasion
417 potential. Therefore, in any such scenario, it is important to take steps to avoid this by
418 synchronising the technology lifetime with the IAP stock.

419

420 One particularly important limitation of the study was that the risk of wildfire on bio-energy
421 from IAPs is uncertain. A wild fire would cause a loss of biomass in the short term, but

422 would densify most of the prominent IAPs in the long term which is regarded as a benefit
423 from a financial perspective. This creates a dilemma from an IAP control perspective and
424 remains a contentious debate. Also, we have found that the current supporting legislative and
425 institutional structures in South Africa are ill equipped to enable the uptake of green energy
426 derived from IAPs. The uptake of decentralised technologies is particularly problematic in a
427 developing country context such as South Africa, mainly because the conventional approach
428 towards energy supply is a centralised one and developments in rural energy supply have
429 often been donor driven, i.e. external agencies determine the ‘suitable technology’ for a
430 particular location. Furthermore, government policy often has contradictions and
431 fundamental differences with regard electricity supply and management. While it is
432 recognised that 100% coverage of dwellings would be physically impossible to achieve
433 because of the migration patterns in deep rural areas, the macro economic policy of the
434 country strongly argue for 100% coverage; yet the dominant electricity supplier (covertly)
435 discourages dispersed generation as do municipalities, because of much needed revenue
436 streams.

437

438 This study suggests that investments in remote and rural energy markets need to be
439 approached in an interactive way to account for the preferences of the target community and
440 to increase the probability of successful uptake. AHP, to a large extent, facilitates this
441 participation process and thus has an important role to play in the successful uptake of policy
442 and management strategies and long-term planning. We have demonstrated and tested the
443 process in the form of a case study on the Agulhas Plain, where different options for utilising
444 IAPs for energy purposes were compared. The rankings as obtained in this study are context-
445 bound, which implies that the findings only have limited application to areas with similar

446 biophysical and socio-economic characteristics. However, the method itself is fully
447 generalisable, and the same prioritisation process can be followed in any study area.

448

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450

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453

454 **4. References**

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675 **Table 1: Estimated biomass available from various species in the Agulhas Plain coastal lowlands based**
676 **on mapped data of these areas (De Lange and Le Maitre, 2010, Kotzé et al., 2010).**

Density class	Total area (ha)	Biomass (t/ha/yr)	Biomass (tonne/yr)
Closed	16414	95	1 559 318
Dense	12638	60	758340
Total	29052		2317658

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695 **Table 2: Summary of the input data variables for the scenarios. Refer to footnotes for explanation of**
696 **technical assumptions and calculations. Data and assumptions from footnotes was sourced from CSIR**
697 **(CSIR, 2010a, CSIR, 2010b)**

Scenario name	Description of technology	Energy balance (considers the energy input required to generate the energy product (excludes distribution) (%)	Description of energy product	Capacity factor (% of time the technology operates at maximum capacity)	Lifetime (years)	Capital outlay (ZAR'000)	Operating cost (ZAR'000/yr) expenditure excludes biomass feedstock supply, marketing and distribution of the energy product as these cost are accounted for separately.	EROI (kWh/ZAR) break even point in brackets (years)
Compressed logs (SC1)	Hammer mill; Drier; Compressed log machine	53 (see note 1)	3854 tonnes of wood-chips are used to produce 3066 tonnes of compressed logs per year (see note 4); which is equivalent to approx. 2 MW energy output per year as used in household stoves for cooking and space-heating.	86	10	ZAR350	ZAR30	6.06 (1)
Slow pyrolysis for charcoal (SC2)	Slow pyrolysis retorts; Briquette machine; Drier	26 (see note 2)	Three retorts use 10950 tonnes wood-chips and produce 4928 tonnes charcoal briquettes per year (see note 5) which is equivalent to approx. 4 MW energy output as charcoal used in household stoves for cooking and space-heating.	86	10	ZAR1650	ZAR330	2.73 (1)
Gasification to electricity (SC3)	Gasification-gas turbine	26	150 KW electricity	75	20	ZAR3000	ZAR150	1.99 (6)
Combustion to electricity (SC4)	Combustion-steam turbine	21	5 MW electricity	75	20	ZAR123000	ZAR2700	1.54 (11)
Decentralised combined heat and power (CHP) (SC5)	Household CHP unit	90 (see note 3)	IAP biomass (4 kg capacity) directly combusted. CHP unit rated at 5 KW. Heat for cooking, hot-water and space-heating. Electricity for 4 LED lights, radio and cell-phone charging.	25 (see note 6)	10	ZAR11	ZAR1	4.74 (1)

698 **Abbreviations:** W- power in **Watts**; J- Energy in **Joules**; g- Mass in **grams**; h- time in **hours**; Multipliers: M- Mega (10^6); K- kilo (10^3);

699 **Note 1:** A 33.5KW mill uses wood chips at 440kg/h (20% moisture @ 18MJ/kg) to produce 350kg/h of compressed logs (5-10% moisture
700 @ 20MJ/kg). Most of these logs are used in a household wood stove which operates at 20% efficiency. The remaining 80% energy is lost as
701 space heat and 40% of this could be recovered as household space-heat. Therefore, the net efficiency is $(60/100)*88=53\%$. Note that this
702 ignores the opportunity to modify the stove to utilise the heat for heating water.

703 **Note 2:** Energy efficiency of 70% (10 tonnes wood @ 18MJ/kg) produces 4.5 tonnes charcoal @ 28 MJ/kg. However, starch (5-10%) and
704 water (25%) are used as binder and must be dried to 5% moisture. As the water specific heat capacity is 4.18 KJ/kg; the latent heat of
705 evaporation of water is 2260 KJ/kg; to raise temperature from 20-80°C : $E1=0.2*4.18*60=50$ KJ. To evaporate the water: $E2=0.2*2260=452$
706 KJ. Therefore, the energy to dry 1 kg charcoal: $E1+E2=0.5MJ$. Since each briquette contain 5-10% starch, some of the energy content of
707 briquettes comes from this starch (starch energy value of 18 MJ/kg) and this is an external energy input. Each kg contains minimum 0.05kg
708 starch equivalent to 0.9MJ. Since charcoal has an energy value of 28MJ/kg the addition of 5% starch and the drying of the briquettes
709 represents an energy cost of $(0.5+0.9)/28$ or 5%. The overall process efficiency is therefore estimated at 65%. The energy balance assumes
710 that charcoal will be used equally in an outdoor grill and indoor household stove. The same assumptions is applied to the household stove as
711 mentioned in Note1.. The outdoor grill is used to cook food at 5% efficiency and 15% of the space heat captured and valued giving net
712 efficiency of 20%. The average overall usage efficiency is $(60+20)/2=40\%$. The energy balance is therefore $0.4*65%=26\%$. Note also that
713 this ignores the opportunity to modify the stove to utilise the heat for heating water.

714 **Note 3:** Energy efficiency is 90% for the stove @ 5KW (2KW for cooking, 2KW hot water geyser and 1KW space-heat). Assumes that <2%
715 of the energy will be used to generate electricity, 38% used to cook, 40% hot water geyser and 20% is space heat. Approximately half of this
716 space heat (ie 10%) will be valuable. Therefore, the net efficiency is 90%.

717 **Note 4:** Assume that 350kg/h compressed logs (@ 5-10% moisture) produced from 440kg/h IAP feedstock (20-30% moisture). The energy
718 content of logs is 20 MJ/kg, so output is equivalent to 1.94MW

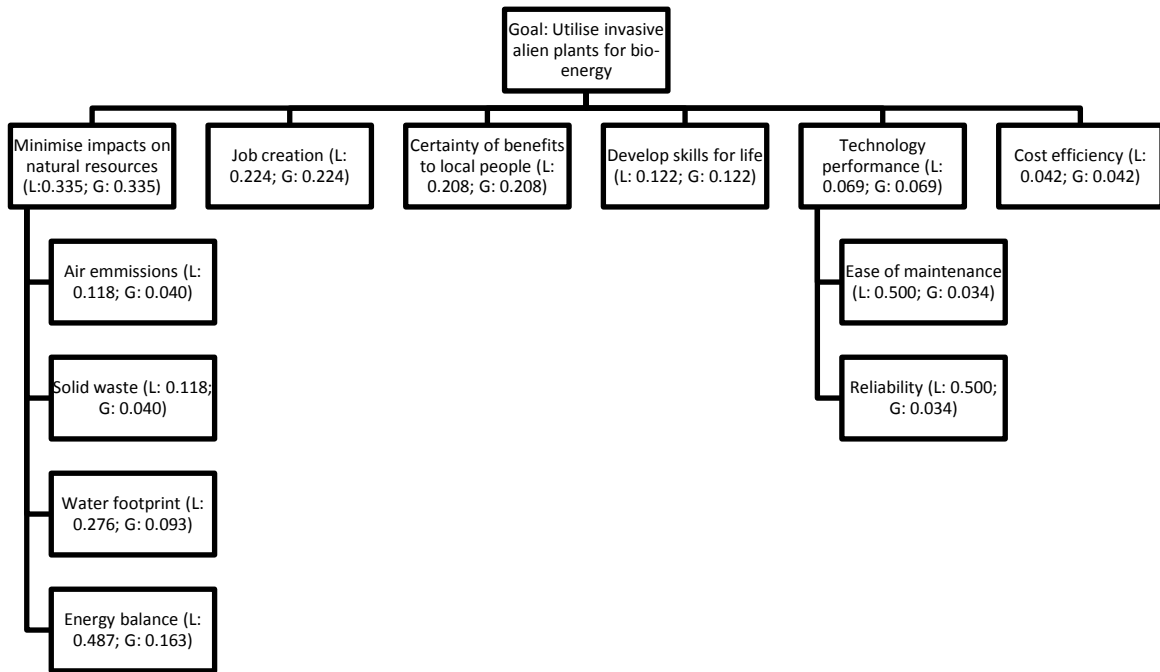
719 **Note 5:** Three retorts utilise 30 tonnes biomass to produce 13.5 tonnes charcoal briquettes per day with an energy content of 28 MJ/kg.
720 Charcoal output is equivalent to 4.34 MW, however 5-10 % of the briquettes is starch binder so the net energy content of the char product is
721 approximately 4MW.

722 **Note 6:** Maximum output would require loading 4kg biomass fuel every 4h (i.e. 6 times per day). However, a 25% capacity factor assumes
723 that the use of the stove will primarily be driven by the need to cook and therefore will be loaded 1.5 times per day (a half load of 2kg in
724 morning and full load of 4 kg in the evening).

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731 **Figure 1: Weighted criteria tree as used in the AHP model. Sub-criteria (“L”) is aggregated to main**
 732 **criteria and the goal (“G”).**

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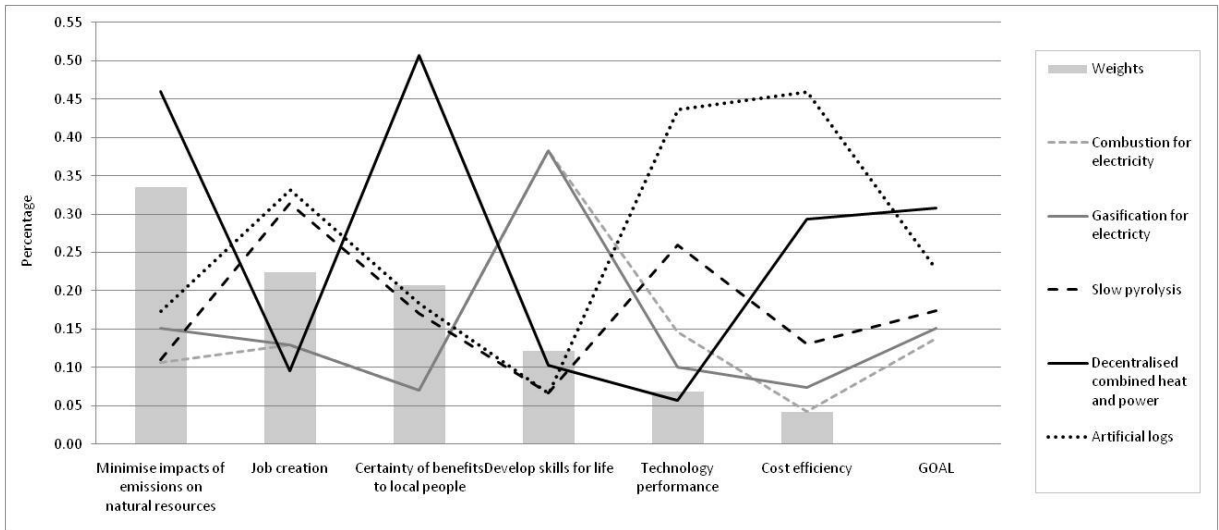
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741 **Figure 2: Results of the AHP**

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