

1 **Drivers of interannual variability in Net Ecosystem Exchange in a semi-arid savanna**
2 **ecosystem, South Africa**

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12 **1. Abstract**

13 Inter-annual variability in primary production and ecosystem respiration was explored using
14 eddy-covariance data at a semi-arid savanna site in the Kruger Park, South Africa. New
15 methods of extrapolating night-time respiration to the entire day and filling gaps in eddy-
16 covariance data in semi-arid systems were developed. Net ecosystem exchange (NEE) in
17 these systems occurs as pulses associated with rainfall events, a pattern not well-represented
18 in current standard gap-filling procedures developed primarily for temperate flux sites. They
19 furthermore do not take into account the decrease in respiration at high soil temperatures. An
20 artificial neural network (ANN) model incorporating these features predicted measured fluxes
21 accurately (MAE 0.42 gC/m²/day), and was able to represent the seasonal patterns of
22 photosynthesis and respiration at the site. The amount of green leaf area (indexed using
23 satellite-derived estimates of fractional interception of photosynthetically active radiation
24 f_{APAR}), and the timing and magnitude of rainfall events, were the two most important
25 predictors used in the ANN model. These drivers were also identified by multiple linear
26 models (MLR), with strong interactive effects. The annual integral of the filled NEE data was
27 found to range from -138 to +155 g C/m²/y over the 5 year eddy covariance measurement
28 period. When applied to a 25 year time series of meteorological data, the ANN model
29 predicts an annual mean NEE of 75 (± 105) g C/m²/y. The main correlates of this inter-annual
30 variability were found to be variation in the amount of absorbed photosynthetically active
31 radiation (APAR), length of the growing season, and number of days in the year when
32 moisture was available in the soil.

33
34 **Keywords:** eddy covariance, gap filling, net ecosystem exchange

35 2. Introduction

36 Carbon dioxide flux measurements using the eddy covariance technique generate a raw
37 dataset with a very high temporal resolution (generally 10-20 Hz). The first step in the
38 analysis of these data is to screen them for spurious values, perform various corrections, and
39 then integrate the fluxes over periods of about 30 minutes. The half-hour data provides
40 important insights into many short-term physiological processes, but most ecological and
41 management-relevant questions are framed over even longer timeframes – from days to
42 years. A matter of particular interest to both ecologists and ecosystem managers is the inter-
43 annual variability of primary production and carbon storage (Lauenroth et al., 2006). Semi-
44 arid savannas are characterised by high inter-annual variability, in response to highly variable
45 rainfall. This underlies many features of their ecology, including the likelihood and intensity
46 of fires, the growth and migration of animal populations, and the stability of the tree-grass
47 mixture (Higgins et al., 2000;Tyson, 1986;Reed et al., 1994;Ma et al., 2007;Serneels et al.,
48 2007), and makes savanna systems particularly hard to manage.

49

50 Accumulating 30 minute flux measurements to longer time periods is not a simple matter of
51 adding them up, for two main reasons. The first is that even the best-run eddy covariance
52 datasets have gaps, due to instrument failure or weather conditions that cause the eddy
53 covariance flux assumptions to be violated. The second is that the eddy covariance
54 measurement, net ecosystem exchange (NEE), is often not what is needed by ecologists who
55 are often more interested in its components, gross primary production (GPP) and ecosystem
56 respiration (R_{eco}):

57 $NEE = GPP + R_{eco}$ (observing the convention that fluxes from the atmosphere to the ground
58 are given a negative sign)

59

60 A model is used to bridge the data gaps in what is intended to be an unbiased fashion. The
61 same or different models can be used to deconvolve the NEE signal into its components. A
62 wide range of standard procedures have been developed for this process, largely for
63 application in temperate ecosystems (Falge et al., 2001; Papale et al., 2006; Moffat et al.,
64 2007). These are not always appropriate for tropical wet-dry systems. They use
65 phenomenological models, neural networks or process-based models to achieve their
66 objectives. The readily-available ones do not work well for data from semi-arid sites in
67 southern Africa. This is because they assume the major controls on flux processes to be solar
68 radiation and temperature, whereas temperatures in the semi-arid tropics are almost always
69 warm enough to permit physiological activity, and insolation is sufficient, at least during non-
70 cloudy days, for light saturation of part or all of the typically-sparse canopy. In arid and semi-
71 arid systems, the main control on the rate and duration of many ecosystem processes is soil
72 moisture.

73

74 As a further complication, in low-rain, high-evaporation ecosystems, where the soils dry out
75 between successive rainfall events (so-called pulse-driven systems), the various terms in the
76 carbon budget are highly dependent on the recent history of the system (Huxman et al.,
77 2004). For example, following a rainfall event, respiration increases rapidly whereas it takes
78 several days for the ecosystem to reach maximum photosynthesis (Williams et al., 2008).
79 Similarly, the magnitude of the system response depends not only on the size of the current
80 rainfall event, but on the amount and timing of preceding events: after a long drought the
81 response to a rain event is larger than to a similar-sized event during the middle of the rainy
82 season, but the time taken to reach the peak response is longer (Veenendaal et al., 2004).
83 Therefore, it is not possible to use instantaneous measures such as the soil moisture content as

84 a sole proxy for the state of the system. Gap-filling therefore requires consideration of indices
85 that have ‘memory’: for instance, accumulators of water deficit.

86

87 Moreover, ‘phenomenological’ models will only be appropriate when they truly represent the
88 underlying responses (Falge et al., 2001). Most current respiration models define the
89 relationship between respiration and temperature using an exponential- or logistic-shaped
90 function; i.e. functions that either continually increase, or level off at a maximum value
91 (Moffat et al., 2007). These models were developed in systems where temperature ranges are
92 generally below 30 °C (Fang and Moncreiff, 2001; Lloyd and Taylor, 1994). Physiologically,
93 respiration is expected to decrease once temperature exceeds the optimum for microbial
94 activity (Yamano and Takahashi, 1983). In tropical dry systems, the soil temperature in the
95 top centimetres often exceeds 40 °C. Thus more appropriate functional forms need to be
96 developed before current gap-filling methodologies can be applied globally.

97

98 Improving functional relationships to include extreme conditions would also be valuable in
99 the context of climate change. In coming decades, many ecosystems around the world are
100 likely to be exposed to higher temperatures and reduced moisture availability. Information on
101 ecosystem responses to high temperatures and intermittent droughts will be valuable in
102 predicting responses to these changes.

103

104 We present a statistical approach to estimating annual NEE for a semi-arid savanna system in
105 southern Africa. We tested the importance of six environmental drivers of daily
106 photosynthesis (GPP) and respiration (R_{eco}) at the Skukuza flux tower in the Kruger Park
107 (25.02° S, 31.50 ° E). Predictors commonly used in temperate systems were included,
108 together with a range of environmental predictors chosen to reflect the effect of pulsed

109 rainfall events. Predictive models were then used to interpolate annual fluxes over a 25 year
110 time period, and to investigate the degree and possible causes of inter-annual variation in CO₂
111 exchange.

112

113 Our approach was motivated by the fact that there was a limited amount and duration of flux
114 data (spanning 6 years with many gaps, which is too short for a reliable estimate of variance),
115 but that a full time series of daily meteorological and phenological data was available for a 25
116 year period. Working at a daily time-step allowed us to bridge the gap between the half-
117 hourly flux data and the crucial annual timescale, and to use the long-term meteorological
118 data to estimate inter-annual variability. Process-based modelling would be ideal for these
119 systems where previous conditions affect the response of the system to perturbation, but we
120 chose to limit ourselves to a statistical analysis, given our imperfect understanding of the
121 processes driving NEE in these systems. Results from this research will be used to develop
122 more process-based models.

123

124 This paper aims to:

- 125 • Document new procedures for eddy covariance gap-filling that are appropriate for
126 dry, hot ecosystems;
- 127 • Explore the factors associated with short-term (daily) variation in NPP, GPP and R_{eco};
- 128 • Calculate annual estimates of NEE and explore the main factors driving inter-annual
129 variation in savanna carbon exchange at the Skukuza flux site in South Africa.

130

131 **3. Methods**

132 **3.1. Study Site**

133 A flux tower situated in a semi-arid savanna near Skukuza, in the Kruger National Park has
134 been collecting data since February 2000. The site is 370 meters above sea level with strongly
135 seasonal rainfall occurring between November and April. Mean annual rainfall is 550 ± 160
136 mm. The landscape is gently undulating, consisting of broad-leaved *Combretum apiculatum*-
137 dominated savanna on the coarse sand crests and fine-leaved *Acacia nigrescens* savanna on
138 sandy clay loam in the valleys (Scholes et al., 2001). The soils are about 0.6m deep. The eddy
139 covariance flux tower is situated at the ecotone between the two vegetation types.

140

141 The woody vegetation reaches 8-10 m in height and the flux sensors are at 17 m, giving the
142 tower a footprint of about 500 m. The vertically projected tree canopy cover in this area is
143 about 30% and woody basal area is $7\text{m}^2\text{ha}^{-1}$. The grass layer is dominated by *Panicum*
144 *maximum*, *Digitaria eriantha*, *Eragrostis rigidior*, and *Pogonarthria squarrosa*.

145

146 The tower is instrumented with a Gill sonic anemometer measuring wind velocity in three
147 dimensions and a LICOR 6262 closed-path infrared gas analyzer measuring water vapour,
148 CO₂ concentration, and pressure. The raw high frequency (10 Hz) data was processed
149 following (Lee et al., 2004) to produce half-hourly measures of above-canopy turbulent
150 fluxes of sensible heat, water vapour, and carbon dioxide. Heat and mass fluxes were
151 calculated based on conventional equations and corrections (see e.g. Moncrieff et al.,
152 1997; Aubinet et al., 2000) and all fluxes are reported as positive upward from the land to the
153 atmosphere. Canopy storage flux was estimated from the half-hourly time derivative of a 16
154 m column integral based on CO₂ concentrations measured at 0.75, 2.0, 3.5, 5.25, and 16 m,

155 and added to the above-canopy turbulent flux for data analysis. Incoming and outgoing long-
156 and shortwave radiation was measured with Kipp and Zonen shortwave and thermal
157 radiometers mounted at 22 m.

158

159 Average half-hourly volumetric soil water content was estimated with 15 cm long Campbell
160 Scientific frequency domain reflectometry probes installed horizontally at soil depths of 3, 7,
161 16, 30, and 50 cm in the clayey *Acacia* –dominated soils downhill of the tower, and 5, 13, 29,
162 and 61 cm in the sandier *Combretum* –dominated soils uphill. Half-hourly averaged soil heat
163 flux was obtained with HFT3 plates (Campbell Scientific) installed 5 cm below the ground both
164 under and between tree canopies. Rainfall per half hour was measured with a tipping bucket
165 rain gauge located on the tower top, along with other standard meteorological variables such
166 as air temperature and humidity, wind speed and direction.

167

168 **3.2. The effect of the ecotone**

169

170 The differences in soil properties and species composition above and below the seepline were
171 expected to be apparent in the flux data from the tower. To test this we separated the half-
172 hourly fluxes into predominantly broad-leafed and predominantly fine-leafed (based on the
173 wind direction) we summarised these monthly over the 6 years (Figure 1). Although night
174 time carbon flux seemed to be slightly higher in the broad-leafed savanna over the dry season
175 (months 4-8) the results were not significant enough to justify running these analyses
176 separately for the two different systems. Kutsch et al (2008 this edition) similarly notes that
177 the data “show no significant differences between the savanna types in terms of fluxes”.
178 Whether this was due to a lack of ability to differentiate between fluxes from the two sites, or
179 because at the landscape level the differences are not significant, we chose to complete the

180 rest of the analysis using all flux data as one unit. We used a model to create an integrated
181 site-level soil moisture record (see Appendix A).

182
183

184 **3.3. Data processing and gap filling**

185 Flux data were available from February 2000 to December 2005 (the site continues to
186 operate, but with an open-path IRGA). Of the half-hourly data, 41% was missing, which is
187 slightly more than the average among flux sites of 35% (Falge et al., 2001). As rainfall
188 occurs during summer months of November to April the flux data were summarised by
189 rainfall years (July to June) which provided five full years of flux data – with data coverage
190 ranging from 30 to 74 % annually. Most of the data gaps were for a single half hour interval,
191 but instrument failure due to lightning strikes resulted in six gaps of over two months
192 duration, usually occurring during summer periods. These large, non-random gaps limit the
193 types of gap filling approaches that can be used.

194

195 When a u^* filter of 0.1 m/s (Reichstein et al., 2005) was applied to eliminate periods of low
196 turbulence during which eddy covariance measurements are unreliable, the missing flux data
197 increased to 49 %. Linear interpolation was used to fill gaps < 2 hours in duration, which
198 reduced the missing data to 44%. These half-hourly data were then summed to calculate
199 daily NEE values for all days with unbroken 30-minute time series. The result was 698 days
200 of NEE data. These days were not randomly distributed through the year, with the rainy
201 months (particularly December and January) represented by much less data than the dry
202 months of June through September (Figure 2). Dry, winter conditions are therefore over-
203 represented in the sample. In addition, one of the periods of most continuous and cleanest
204 observations spans an intense drought, 2002-2003 growing season, further biasing results.

205

206 Simple gap-filling techniques using mean daily averages are inadequate for filling gaps in the
207 Skukuza data because the stochastic and variable NEE response over the course of a wetting
208 event would not be well represented by a summary value, and because gaps in the data often
209 span several weeks. Non-linear regression methods work well when there is just one main
210 driver of carbon uptake or release (in temperate systems, temperature is normally used to
211 drive respiration, and APAR to drive photosynthesis; (Moffat et al., 2007). However, the
212 presence of multiple drivers at the Skukuza site means that single-parameter non-linear
213 methods are unlikely to be sufficient.

214

215 Similarly, Marginal Distribution Sampling (MDS: Reichstein et al., 2005) fills gaps by taking
216 the average value for data collected under similar meteorological conditions within a certain
217 window of the missing data. In this way it accounts for temporal auto-correlation as well as
218 co-variation with meteorological drivers. However, when there are long gaps this method
219 breaks down, and the choice of “similar” meteorological conditions requires that the
220 appropriate hydrological are considered in the model – current implementations use only Rn,
221 temperature, and VPD.

222

223 We used Artificial Neural networks (ANN) as our gap-filling approach, as this method
224 accommodates non-linear relationships between variables but requires few *a priori*
225 assumptions on the relative importance of different variables or their functional relationships.
226 The usefulness of ANNs depends entirely on the appropriate selection of input variables –
227 and we hoped to improve on standard methods available by choosing variables which would
228 reflect the pulsed response to soil moisture in arid systems. We also ran standard multiple
229 linear regression models on the data to explore interactive effects between the variables. This

230 approach allowed us to investigate the important drivers of NEE, as well as develop models
231 which could be used for prediction using long-term meteorological data.

232 **3.4. NEE, Photosynthesis, Respiration**

233 Half-hourly night-time fluxes were used to estimate the day-time respiration. A stricter u^*
234 threshold of 0.25 ms^{-1} (Kutsch et al., 2008) was used for this analysis, as it was more
235 important to have reliable data than large sample sizes. Respiration is controlled by
236 temperature, which generally varies quite predictably over the course of a day, as well as
237 variables such as soil water content and the amount of actively photosynthesising leaf
238 material, which are relatively constant over a single day, but vary over longer time scales. We
239 therefore took a two-scale approach to determining day-time ecosystem respiration: we
240 derived a temperature response curve by fitting it to ‘optimum’ respiration conditions – i.e.
241 the maximum values measured at a range of temperatures (all valid half-hourly night-time
242 fluxes were used for this). This curve was used to estimate the maximum potential respiration
243 rate for each daylight interval, using the daytime temperature trend as input (see appendix B
244 for more details on this method). The actual respiration during any particular day was then
245 estimated as the temperature-driven ‘potential’ scaled by the ratio of observed night-time
246 respiration to the potential night-time respiration for that day. This scaling factor was
247 assumed to account for the effects of soil moisture and physiological activity. Unlike the
248 flux-partitioning method of Reichstein et al (2005) this method does not require a separate
249 temperature response function to be derived for each day.

250

251 Conventional Arrhenius or Lloyd-Taylor temperature functions were not considered
252 appropriate representations of the response functions, as day-time temperatures at the site
253 often exceed that which is optimum for microbial activity (Yamano and Takahashi, 1983).
254 An analysis of independently-collected respiration data from the site, collected using soil

255 chambers, indicated that a generalised Poisson temperature relationship produced the best fit
256 to measurements of soil respiration (Kirton et al in prep).

257

258 We therefore used the following equation to describe the optimal temperature response:

259

$$260 \text{ Respiration} = M \left(\frac{b - \text{Soil temperature}}{b - a} \right)^c \exp \left\{ \left(\frac{c}{d} \right) \left[1 - \left(\frac{b - \text{Soil temperature}}{b - a} \right)^d \right] \right\}$$

261

262 Parameters were estimated using a non-linear least squares by means of the Levenberg-

263 Marquardt algorithm:

$$264 \hat{M} = 1.0104(0.0814); \hat{a} = 27.6815 (1.5876); \hat{b} = 11.4221(10.2024); \hat{c} = 0.6676 (1.1693); \\ \hat{d} = 4.1457(1.7782)$$

265 where values in brackets represent the standard error of the estimate. Only days when there
266 were more than three valid night time flux values with which to estimate the scaling
267 parameter were used to interpolate day-time fluxes. See Appendix B for details on this
268 method and a comparison with other methods.

269

270 Negative night time fluxes were excluded from the model fitting, as there was no theoretical
271 justification for negative respiration. Interpolated respiration values that dropped below zero
272 (which can occur at very high or low temperatures, using the parabolic curve) were given a
273 value of zero. This method produces predicted respiration values with similar distributions to
274 those recorded for all conditions of soil moisture and f_{APAR} (Figure 3).

275

276 Daily respiration (R_{eco}) values were obtained by calculating a half-hourly value (multiplying
277 the per second value by 60×30) and summing this over the 48 half-hours. All other daily

278 values were calculated in the same way. Daily Gross Primary Production (GPP) was
279 calculated by subtracting the interpolated day-time respiration values from the recorded
280 daytime NEE values, and summing over the daylight hours. This resulted in a dataset with
281 372 valid records for R_{eco} and 529 for GPP.

282 **3.5. Drivers of NEE**

283 In temperate systems incoming solar radiation (PAR) and temperature are the main drivers
284 used to predict photosynthesis and respiration. In some models these are modified by
285 measures of LAI and soil moisture (Moffat et al., 2007). We chose to test six input variables
286 as predictors of GPP and R_{eco} (see Table 1).

287

288 Only data that could be derived from standard daily South African Weather Services (SAWS)
289 climate records or long-term low-resolution satellite vegetation indices were used as input
290 predictors, in order that the models could be used in conjunction with the long-term records
291 to estimate NEE over periods much longer than the eddy covariance data would permit. The
292 daily time-course of temperature variables was estimated from daily maximum and minimum
293 air temperature. Soil water content was modelled using a simple bucket model and Penman-
294 Monteith evapo-transpiration functions (Archibald and Scholes, 2007). The half-hourly
295 meteorological data available at the flux tower was used to validate these models (see
296 Appendix A)

297

298 Three different measures were used to indicate the hydrological state and history of the
299 ecosystem: Relative Plant Available Water (RPAW); water deficit (a function which
300 accumulates the deficit for all days of water stress $\theta < \theta_{crit}$ until rewetting occurs); and time
301 since wetting (the time since the last big wetting event – i.e. time since θ increased above
302 θ_{crit}). Equations for these indices can be found in Table 1. Mean air temperature – which

303 correlates well with soil temperature (Appendix A) - was used as the predictor of R_{eco} ,
304 whereas mean daytime temperature was used as the predictor for GPP. The European Joint
305 Research Centre 10-day f_{APAR} product (Pinty et al., 2002) was linearly interpolated to create
306 a daily f_{APAR} parameter. A relationship between AVHRR-derived NDVI (the 'GIMMS data',
307 (Tucker et al., 2005)) and f_{APAR} was used to define the daily f_{APAR} input for the period before
308 the beginning of the Joint Research Centre (JRC) dataset (See Appendix A).

309 **3.6. Modelling approach.**

310 Two different artificial neural network (ANN) methods were tested: Generalised Regression
311 Neural Network (GRNN) and Multi-Layer Feed Forward Neural Network (MLF). The
312 GRNN is based on a kernel smoothing approach and has the advantage of using non-
313 parametric regression procedures (which makes no assumptions about the underlying data)
314 and can be trained quickly as only the smoothing parameter needs to be estimated and
315 optimised. As has been found in other studies (Cigizoglu, 2005;Currit, 2002;Kisi, 2006) this
316 method is efficient for modelling non-linear systems and worked as well as the more
317 traditional MLF, which required excessive fine-tuning to optimise the system architecture.
318 Three separate models were developed for predicting R_{eco} , GPP, as well as daily NEE.
319 Models were developed using 80% of the data for training and 20% for testing (proportions
320 of 70-30% were also tried, without substantially changing the results).

321
322 Multiple linear regression equations with up to three-way interactions were examined for
323 both photosynthesis and respiration. A combination of backward selection and stepwise
324 selection was used to obtain significant predictors in the model. The ability of the MLR to
325 explore the importance of different variables separately and in combination added value to
326 the results of the ANN. However, there are strong theoretical reasons against using ordinary
327 least squares (OLS) regression for data-filling (Richardson and Hollinger, 2005), which is

328 why we restricted their use to exploring the relationships between variables. Many of the
329 meteorological variables, at least over a certain range, are expected to have a near-linear
330 relationship with respiration and photosynthesis. Temperature is an exception: therefore
331 quadratic terms of temperature were also included during the model selection process.

332

333 **3.7. Error estimation**

334

335 The random error component of the total error in the daily carbon fluxes was considered in an
336 attempt to obtain a confidence interval for the annual estimates of NEE. The systematic
337 component of the error was not assessed for this paper, but this analysis will be carried out at
338 a later stage. To estimate the random error, the method described by Richardson et al (2008)
339 was used, where the model error was used as a surrogate for the random error. The error of
340 the daily ANN model prediction (difference between the observed and modelled daily fluxes)
341 was calculated for all cases where the observed daily fluxes were available. The distribution
342 of these errors fitted a Laplace distribution better than a normal distribution (Chi-squared
343 tests for goodness of fit were $\chi^2 = 37.37$ compared with $\chi^2 = 111.01$ for the normal
344 distribution). Richardson *et al.* (2008) also found the errors in half-hourly flux data to be
345 distributed according to the Laplace distribution.

346

347 We assumed the daily errors were independent and identically distributed. This allowed us to
348 use the Central Limit Theorem to assume normality for the annual sum of the errors in the
349 fluxes. The expected value of the errors was assumed to be zero and the variance of the errors
350 was estimated by the sample variance. The approximate standard error for the annual
351 estimates was then calculated to be $12.9 \text{ g C m}^{-2} \text{ year}^{-1}$, and therefore the error in the annual
352 NEE estimates is $25.3 \text{ g C m}^{-2} \text{ year}^{-1}$ with 95% confidence. This agrees with the estimate of

353 random error obtained by Richardson and Hollinger (2005), where they used the Monte Carlo
354 simulation to estimate the error in the model parameters and model estimates.

355 **4. Results and Discussion**

356 **4.1. Carbon Balance**

357 The diurnal time-course of NEE is highly responsive to soil moisture and the presence of
358 green leaves (Figure 4). Interestingly, maximum CO₂ uptake occurs during periods of low
359 soil moisture when green leaves are still present (Williams et al, 2008), because under these
360 circumstances the contribution of soil respiration is low, but a substantial amount of
361 photosynthesis is still occurring using water stored in the plant, or accessed from deeper soil
362 layers that do not contribute much to ecosystem respiration.

363

364 **4.2. Gap-filling: Modelling R_{eco} and GPP**

365 Despite the relative paucity of daily data both the ANN and multiple regression methods
366 produced models which reasonably represented the input data (Table 2). Mean absolute error
367 (MAE) ranged from 0.37 to 0.56 g C/m²/day, which compares favourably to the 1-1.5 g
368 C/m²/day range of values reported by Moffat et al (2007) for a range of gap-filling methods
369 and vegetation types. Respiration was generally harder to predict than photosynthesis, and the
370 linear models performed badly in predicting R_{eco} (r^2 of 0.41, MAE of 0.68 g/m²/day).

371

372 The ANN identified available green leaf material (indexed by f_{APAR}) to be the most important
373 predictor of both R_{eco} and GPP, but f_{APAR} was relatively more important for predicting GPP
374 than for predicting R_{eco} , as would be expected (Table 3). We interpret the role of f_{APAR} in
375 driving R_{eco} as reflecting the availability of readily-respired substrate. For GPP the time

376 since wetting event was the next most important predictor, which corroborates findings of
377 Williams et al (2008) that there is a delay in the pulse of photosynthetic activity after a
378 rainfall event. In terms of water relations, soil moisture content was the best predictor for
379 R_{eco} , but water deficit and time since wetting were also identified as important. Interestingly,
380 temperature did not prove to be important in predicting either respiration or photosynthesis.
381 This could reflect the daily time-step at which we did the analysis – in this sub-tropical
382 system temperature variation between days and over the growth season is much less
383 important than variation in leaf dynamics and soil moisture in driving NEE.

384

385 For respiration models using Multiple Linear Regression, f_{APAR} and time since wetting were
386 the most significant single predictors. Interactions between various soil moisture parameters
387 and f_{APAR} also significantly improved the fit of the respiration model. As can be seen in
388 Figure 4, the effect of a parameter like soil moisture greatly depends on the amount of
389 photosynthesising green leaf material, so it is unsurprising that these interaction terms are
390 important.

391

392 In the photosynthesis model soil moisture was very significant, and three-way interactions
393 between f_{APAR} , soil moisture, PAR, and time since wetting were important in improving
394 model fit. The importance of the interactive terms perhaps goes some way to representing the
395 delayed photosynthetic response to wetting events identified by Williams et al (2008). It
396 usually takes 5-7 days in this system before photosynthesis reaches its maximum after a
397 wetting event, and this response depends on how much leaf material is present. Temperature
398 was included in both the GPP and R_{eco} models as it produced significant interactions with
399 other variables, but as a main effect it was not significant. The ANN net ecosystem exchange
400 model had the lowest error (Table 2), so this model was used to gap-fill the 6 year dataset .

401 **4.3. Inter-annual variability**

402 Annually-integrated net ecosystem exchange varied from -138 to +155 gC/m²/y over the 5
403 year period for which there was flux data (Table 5). In drought years limited carbon uptake
404 occurs even during the height of summer, but in years with above average rainfall the site can
405 be a sink of carbon for several months of the year (Figure 5). Only two of the five years had
406 negative NEE (in other words, were net carbon sinks at the annual timescale). It is possible
407 that our gap filling methods over-estimate the amount of respiration occurring at this site:
408 there was very little data available during the summer months (Figure 2), so the model was
409 probably not well trained to identify days of maximum GPP in this system. To test this we
410 will need to acquire a more extensive summer dataset for this site. Estimates of random error
411 suggest that years where predicted annual NEE was within +-20 gC/m²/y might actually have
412 been close to carbon-neutral.

413
414 When the 25 year NEE sequence is predicted the pattern becomes more obvious (Figure 6).
415 The site was predicted to be a net sink for carbon in only 6 of the 25 years, but three other
416 years (1989, 1996, and 2000) may have been near-sinks. The data give a long-term mean
417 annual NEE of 75 (±105) g C/m²/y. Loss of a cohort of aging *Acacia nilotica* trees at the site,
418 and increased stem damage with increasing elephant populations over the last 20 years might
419 both contribute to making this site appear as a net sink in this analysis. Recent field data at
420 the site record high rates of tree turnover - 8 % (+- 3%) per annum – with damage by
421 elephants and senescence of old *Acacia nilotica* trees being the main cause (Archibald –
422 unpublished data). These turnover rates are high, but not exceptional for southern African
423 savannas (Shackleton, 1997), and it is perfectly feasible that tree growth could match these
424 losses. Therefore, it would be precipitous to speculate further on the implications of the long-

425 term predictions until there is better information on tree productivity, and more peak-growing
426 season flux data to calibrate the models with.

427

428 The most useful information provided by the long-term prediction is estimates of the inter-
429 annual variation for this site. Figure 7a indicates that there is a strong relationship between
430 predicted annual NEE and available photosynthetically active radiation (APAR, which is
431 $PAR * f_{APAR}$). This analysis suggests that once annually accumulated APAR exceeds about
432 675 MJ/m^2 , the system becomes a sink for carbon (Figure 7a).

433

434 It might seem surprising that soil moisture, which was so important at a daily time scale, does
435 not show a stronger relationship with annual NEE. Even when photosynthesis and respiration
436 are considered separately (Figure 7 b,c), by far the best relationship is found with APAR.

437 This result makes sense when one considers that both the ANN and the MLR analyses
438 showed strong interactive effects of soil moisture with fapar – i.e. the effect of available soil
439 moisture in driving Pn and Re depends heavily on the amount of photosynthetically active
440 green leaf material. Similarly, soil moisture has been shown to be an important driver of
441 seasonal patterns of leaf display at the site (Archibald and Scholes 2007). Fapar can therefore
442 be seen as an integrated measure of hydrological conditions at the site, which is better at
443 predicting annual-scale carbon exchange than any measure derived from short-term
444 measurements of daily soil moisture. For example in the 2003-2004 rainfall year the total
445 annual rainfall was above average (618mm) but it was heavily skewed towards the last part of
446 the growing season. In this instance integrated values of APAR would represent the growing
447 conditions for a season better than total rainfall, or even number of growing season days.

448

4.4. Other pathways of carbon loss from the system

449 A savanna carbon budget would be incomplete without a consideration of fire and herbivory.
450 The fluxes of CO₂ to the atmosphere via these two pathways have not been directly measured
451 at the Skukuza site, but can be inferred and constrained from other data. The abundant large
452 mammalian herbivore (>5 kg body mass) community in this landscape consists of 14 species,
453 mostly Bovidae. The combined herbivore biomass is 3155 kg km⁻² (Scholes et al., 2004).
454 Taking into account the effect of body mass on metabolic requirements and digestability, this
455 translates to a herbivore respiratory flux of 4.5 g Cm⁻²y⁻¹ and a flux from the decomposition
456 of dung of 5.0 g Cm⁻²y⁻¹. The uncertainty range associated with these estimates is unknown,
457 but thought to be around 20%, related mostly to errors in game census. The inter-annual
458 variability is thought to be relatively low. The herbivore respiration and dung decomposition
459 fluxes are subsumed in the ecosystem respiration measured by the eddy covariance system
460 (Table 6).

462
463 The mean fire return time in this landscape in the KNP is 4.2 years (Van Wilgen et al.,
464 2000). The most comprehensive set of fuel measurements for this landscape was taken in
465 August 1992 at 10 locations within 30 km of the Skukuza site (Shea et al., 1996). The
466 combusted material was predominantly dry grass (1442 ± 975 SD kg ha⁻¹), tree litter (1452 ±
467 636 kg ha⁻¹) and a contribution from dead wood (226 ± 194 kg ha⁻¹) giving a total of 3120 ±
468 1795 kg ha⁻¹. A multi-site, multi-year mean grass fuel load for the KNP is 3359 kg ha⁻¹, with
469 a range of 1152-6728 (Trollope and Potgieter, 1985). The emission factor for CO₂, measured
470 for the same fires as the above fuel loads (Ward et al., 1996) is 1699 ± 33 gCO₂ kgDM⁻¹.
471 Therefore, the long-term annualised emission of CO₂ through fire is around 136 ± 58 gCO₂
472 m⁻²y⁻¹. An additional 6.4 ± 3.9 gCO m⁻²y⁻¹ and 0.2 ± 0.2 g CH₄ m⁻²y⁻¹ are also emitted from
473 fires, so the total pyrogenic carbon losses are around 40.0 ± 17.5 gCm⁻²y⁻¹ (Table 6).

474

475 The flux site has burned five times since 2000¹, which suggests that the pyrogenic emissions
476 during this period are probably about twice the long-term, landscape-scale averages
477 calculated above. The pyrogenic fluxes are in principle part of ecosystem respiration, but in
478 practice are not measured by the eddy covariance system because they occur briefly, and
479 during that period exceed the measurement range of the infra-red gas analyser. The inter-
480 annual variability is high because a given site does not burn at all in most years, and the fuel
481 load varies greatly in the years when it does burn, in response to the variability of rainfall in
482 the preceding season.

483 **5. Conclusions**

484 Inter-annual variability in carbon exchange at the Skukuza flux site is on the same scale as an
485 oak savanna in California (Ma et al., 2007). The variability seems to be largely controlled by
486 variations in the length of time that green leaf is displayed by the trees and grasses, and by
487 changes in seasonal patterns of water availability (Figure 7) – both ultimately driven by
488 variations in rainfall between years.

489

490 The flux-partitioning and gap-filling procedures developed in this paper are a distinct
491 improvement on standard methodologies largely because they use more appropriate
492 temperature-response functions and explicitly include a soil moisture control, including
493 indices of the wetting history. Estimates of annual CO₂ flux obtained through gap-filling
494 using an ANN may be slight over-estimates (i.e., slightly biased toward the sink side),
495 because of the paucity of peak growing season flux data. However, it is also possible that this
496 particular savanna site has been a carbon source in recent years due to high tree turnover.

¹ Aug 2000, Aug 2001, Apr 2005, Nov 2006, May 2007

497 Results of the ANN gap-filling procedures and MLR models indicate a large degree of
498 interaction between driver variables and lend support for the development of a process-driven
499 model for this system. Such a model would need to include explicit measures of leaf mass,
500 soil moisture and temperature.

501

502 The generalised Poisson function used here to fit an optimum temperature response curve is
503 an effective method for extrapolating day-time respiration in systems where temperatures
504 often exceed 30°C – provided a scaling factor is used to control for the co-limiting factors of
505 LAI and soil moisture. At a daily to seasonal level, however, temperature was shown to be
506 less important than other factors in influencing NEE.

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513 Environment Operating Unit.

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- 628
- 629

630 8. Figure legends

631

632 Figure 1: Showing monthly carbon (F_c) and water (LE) flux at the site for the two main
633 vegetation types (calculated from the wind directions). Carbon fluxes are separated into
634 daytime and nighttime fluxes.

635

636 Figure 2: Seasonal distribution of valid NEE data points from a six-year long dataset at the Skukuza
637 flux tower.

638

639 Figure 3: Distribution of observed (black) and interpolated (red) half-hourly respiration
640 values over temperature. Data are presented for all conditions, for periods of low soil
641 moisture, for periods with little leaf material (low f_{APAR}), and for conditions of low soil
642 moisture and f_{APAR} . Interpolated values lie well within the distribution of observed values for
643 all conditions. It is also clear that respiration drops off at high temperatures, and that
644 temperature-response functions need to include this reduction at high temperatures if they are
645 to be appropriate for this site.

646

647 Figure 4: Daily time-course of NEE averaged over 5 years of measurements and for six
648 combinations of environmental conditions at the Skukuza flux site. Maximum CO_2
649 sequestration occurs when soil moisture is low but green leaves are still present. Wet
650 conditions were defined as periods when the soil moisture was greater than 9 % volumetric
651 water content, dry conditions, less than 6 %. Periods with green leaves were defined as
652 periods when the f_{APAR} value was greater than 0.2. The average number of days each year for
653 each combination of physiological and soil moisture conditions are shown, together with the
654 average daily sum of NEE ($gC/m^2/day$) for these conditions.

655

656 Figure 5: Annual time course of NEE for two consecutive years (a dry year and a near
657 average year) at the Skukuza flux tower. Red line represents measured daily NEE, blue is
658 modelled using an artificial neural network and inputs of f_{APAR} , soil moisture, temperature,
659 time since wetting, and water deficit.

660

661 Figure 6: Annual NEE estimated over a 25 year time-series at the Skukuza Flux site. Bars
662 represent estimated annual sum, lines show 95% confidence based on random error
663 estimation.

664

665 Figure 7: Relationship between annual NEE (a) R_{eco} (b) and GPP (c) and four potential
666 drivers of inter-annual variability in carbon uptake: annual rainfall, available
667 photosynthetically active radiation, length of the growing season, and number of growth
668 days. Annual rainfall seems to be the least significant, compared with parameters that include
669 seasonal variation in leaf display (APAR and length of growing season), and the seasonal
670 distribution of rainfall. Solid circles represent years 2000-2005 for which flux data were
671 available to constrain the model.

672

673 Figures for Appendix B:

674 Figure B1: Showing the six temperature response functions fitted to the half-hourly night
675 time fluxes (respiration). Plot A shows the parabolic functions fitted over the manually
676 selected maximum points (top function), the automatically selected maximum points (middle
677 function) and the manually selected top of the data mass (bottom function). Plot B shows the
678 Generalised Poisson function fitted over the same three selections of points.

679

680 Figure B2: Showing the distribution of the respiration data interpolated using six different
681 methods (solid points: median values, box: $\pm 25\%$ quantiles, bar: data range). The median
682 and $\pm 25\%$ quantiles are very similar for each method, but the method that calculates the
683 fitted values had slightly lower maxima than the other two methods. All data are well within
684 the range of measured R_e values (u^* -corrected half-hourly night-time fluxes).

685

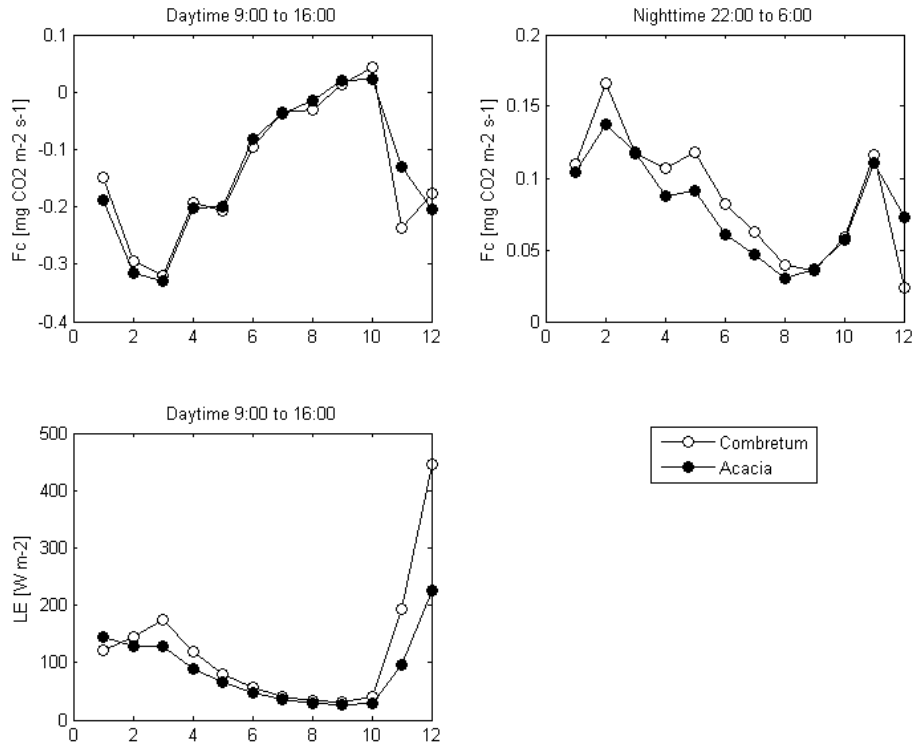
686 Figure B3: The distribution of measured half-hourly night-time fluxes (black circles) and
687 interpolated half-hourly respiration (red crosses) along a temperature axis. Interpolated fluxes
688 represent all half-hour values which had soil temperature data and at least three night-time
689 fluxes to estimate the scaling parameter.

690

691

692 **9. Figures**

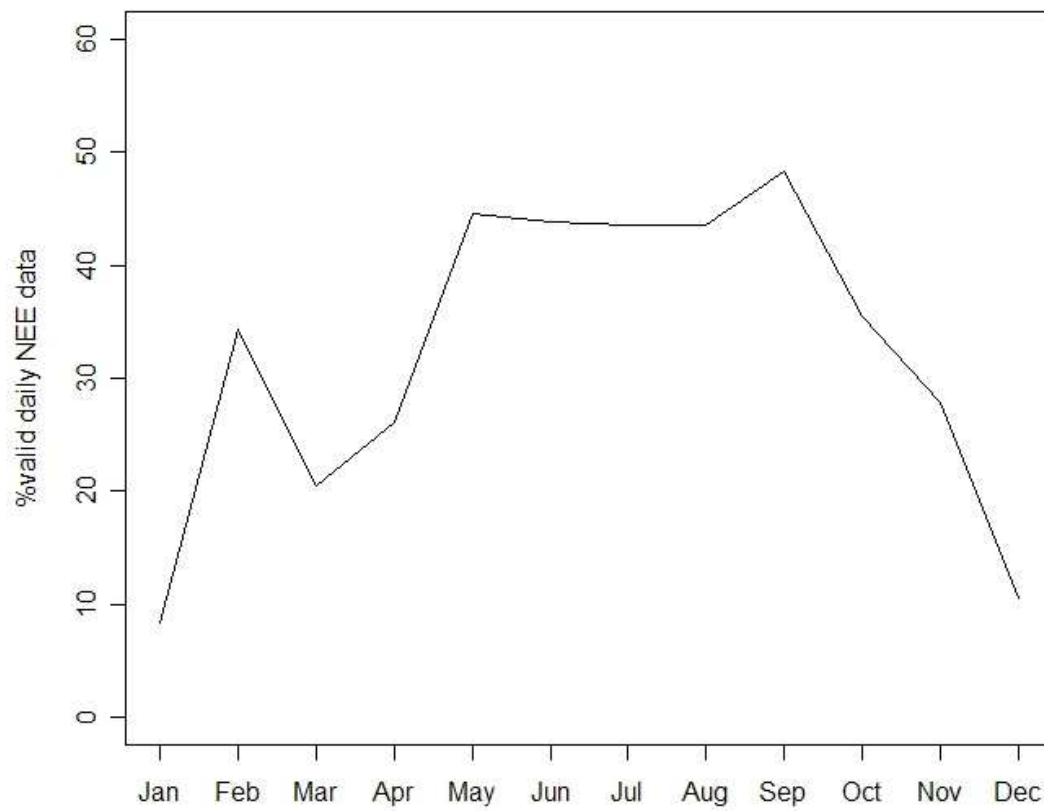
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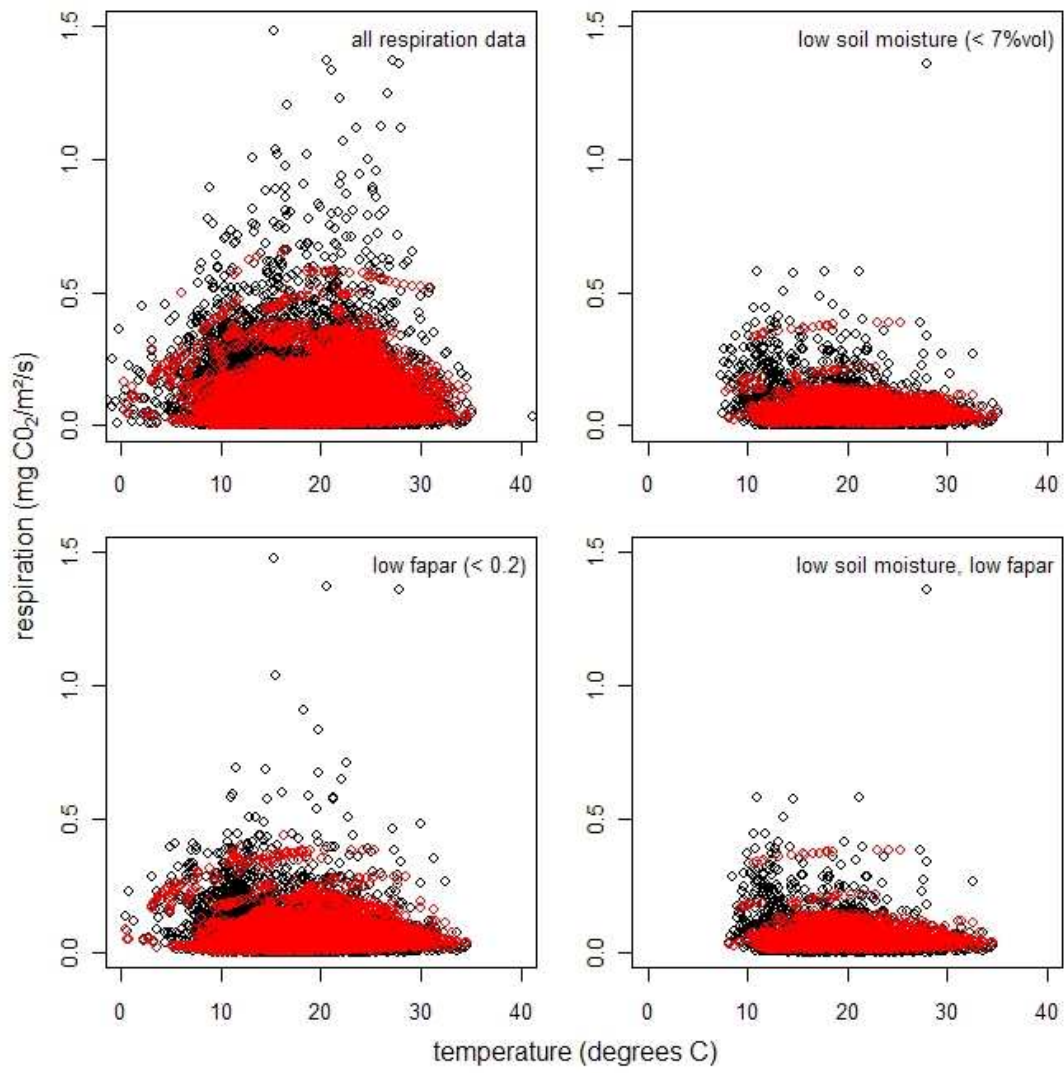
696 Figure 1



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699 Figure 2



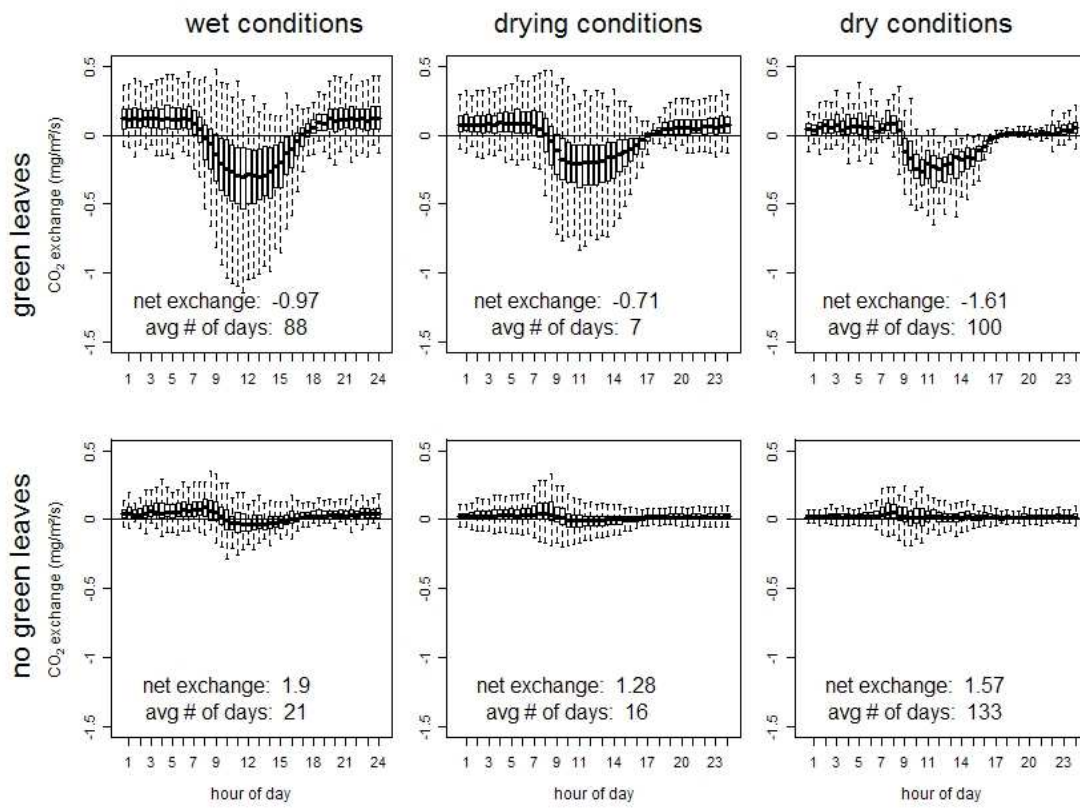
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701 Figure 3

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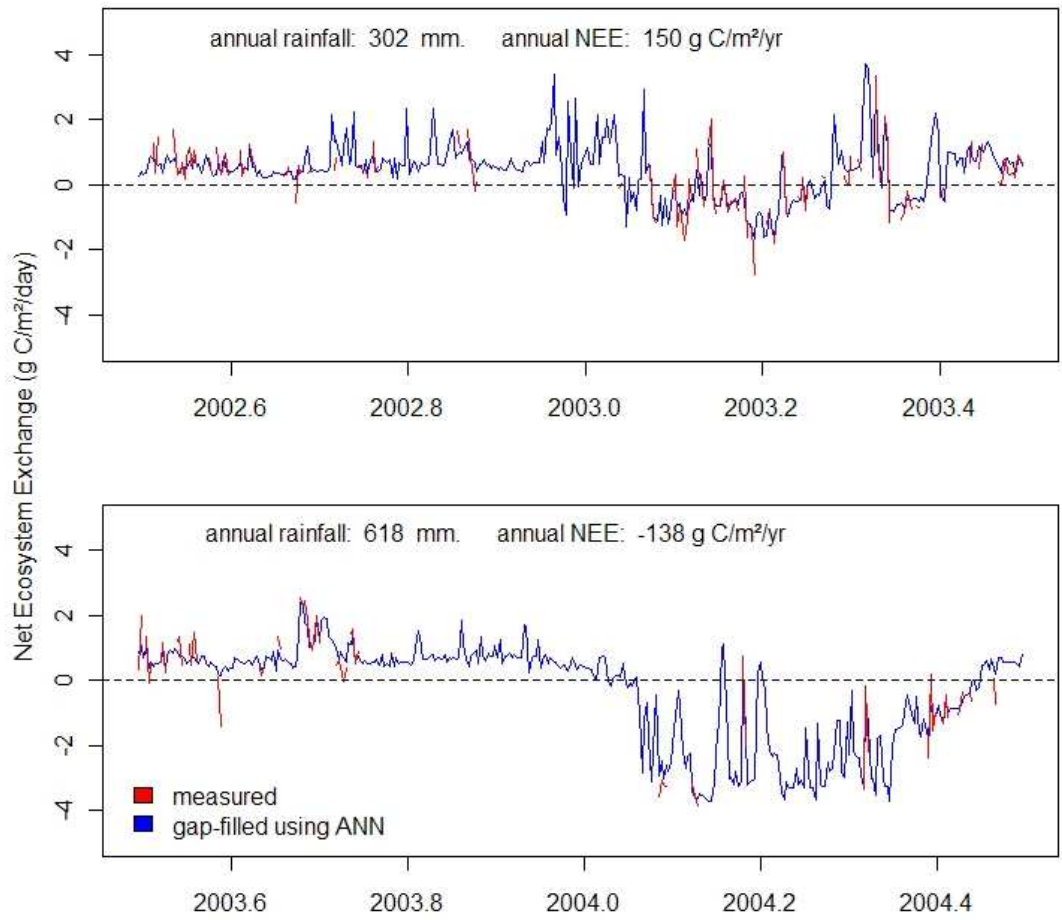
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707 Figure 4



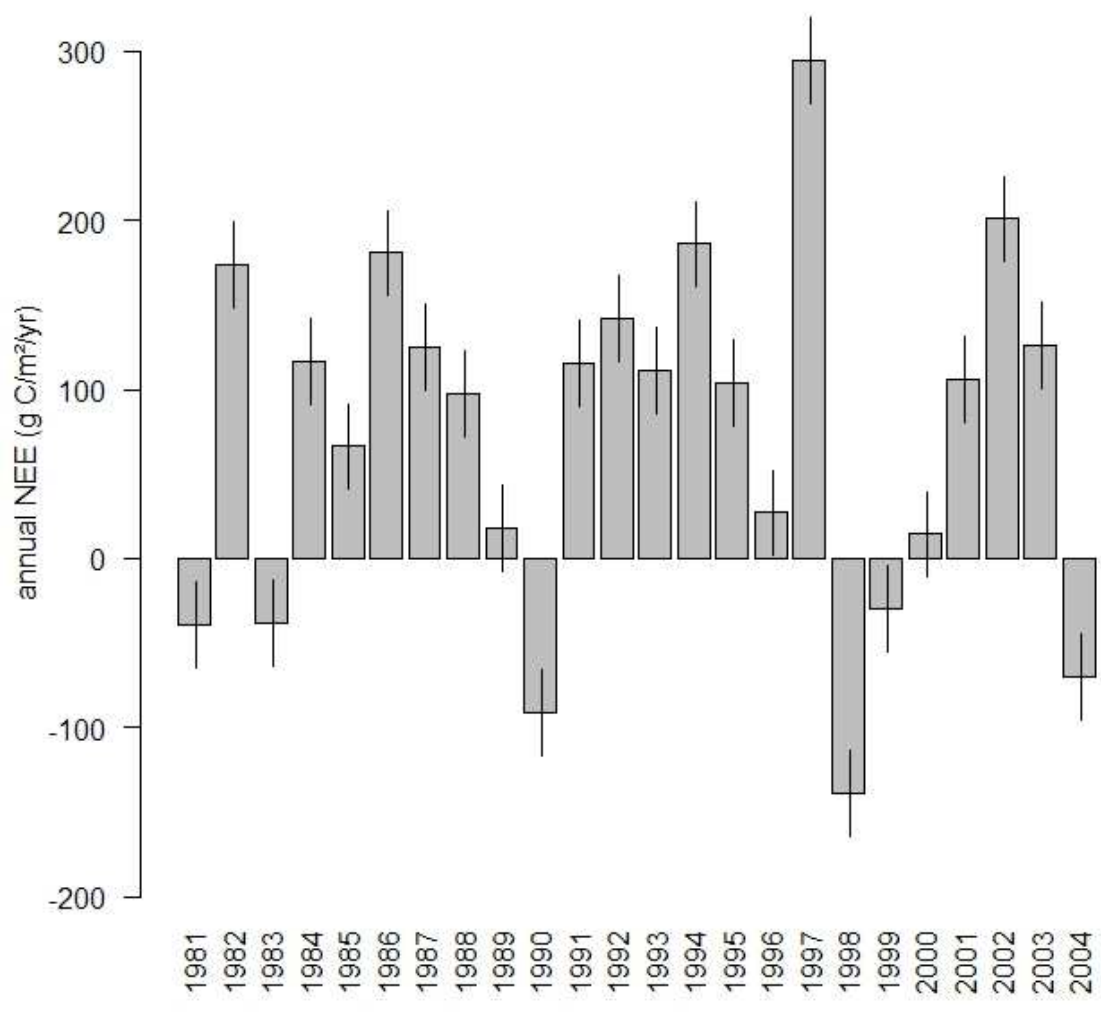
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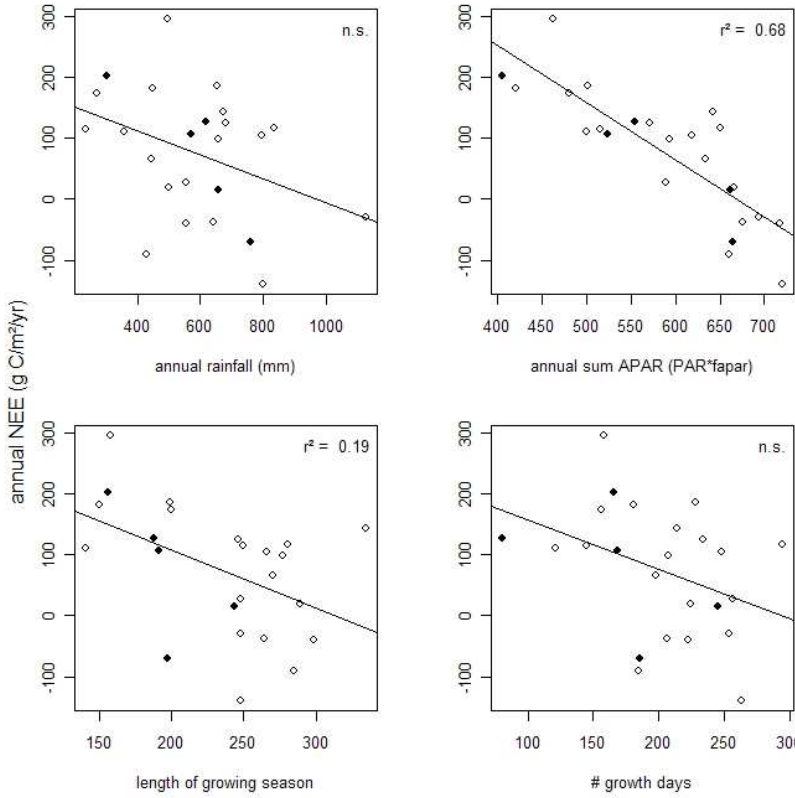
711 Figure 5

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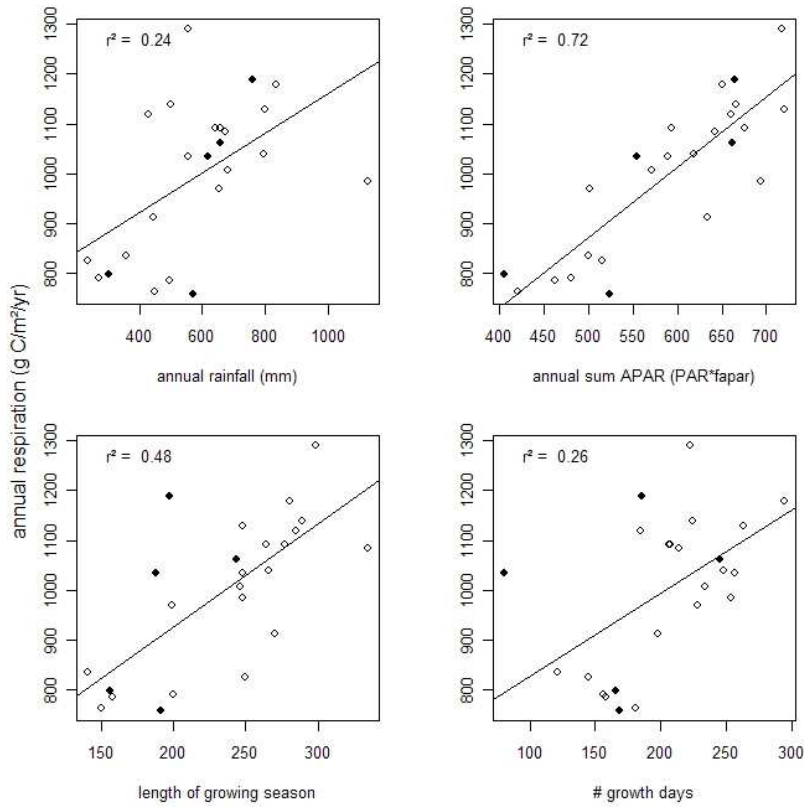
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Figure 6



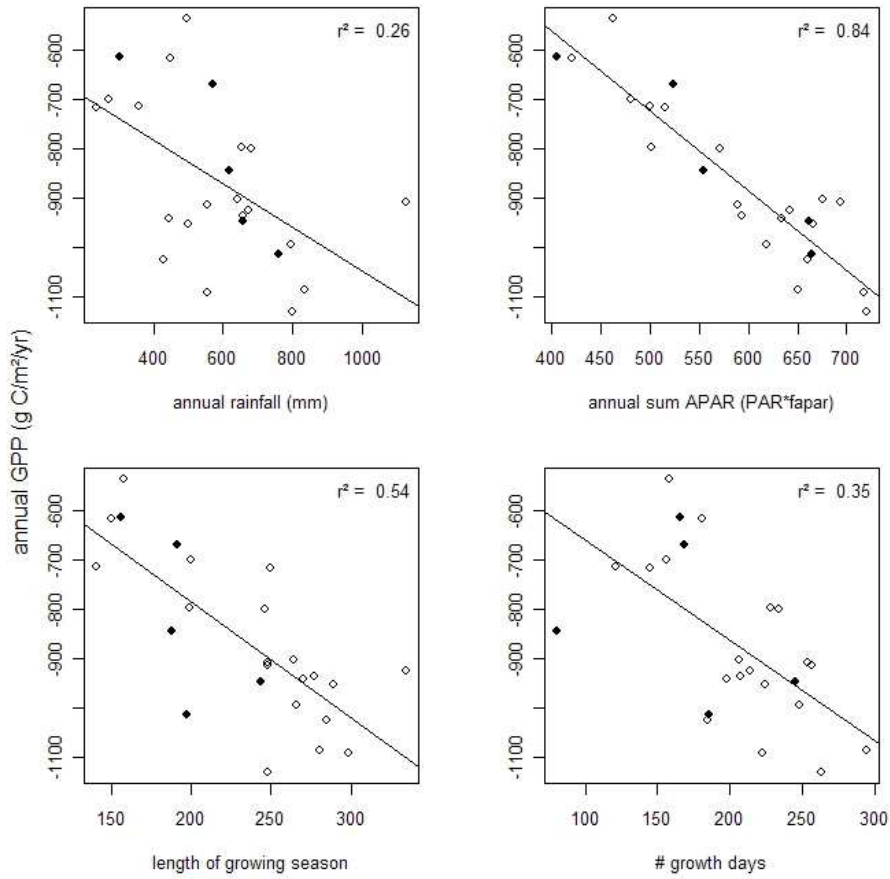
720 a)

721



722 b)

723



724 c)

725

726 Figure 7

727

728 TABLES:

729 Table 1: Defining the six input variables used in the models to predict GPP and R_{eco} . All input
 730 variables were derived from data available at a daily level from the SA Weather Services, so
 731 they could be used to produce long-term predictions
 732

Parameter		Derivation	GPP predictor	R_{eco} predictor
Photosynthetically Active Radiation	PAR	Modelled (energy balance)	X	X
Mean temperature during the day	T_{pn}	$T_{min} + 0.75*(T_{max}-T_{min})$	X	
Soil temperature	T_{re}	$(T_{max} + T_{min})/2$		X
Fraction of absorbed PAR	f_{APAR}	Modelled from satellite-derived reflectances (JRC: http:// fapar.jrc.it/Home.php)	X	X
Relative Available Water Content (RAWC)	θ_{rel}	$(\theta-WP)/(FC-WP)*100$	X	X
Accumulated water deficit	water deficit	If($\theta < \theta_{crit}$) $\Sigma (\theta_i - \theta_{crit})$ If($\theta > \theta_{crit}$) 0	X	X
Period of wet soils	time since wetting	(While $w_{def} = 0$) Σ days since $w_{def} = 0$	X	X

733

734 Table 2: Comparison of model performance. Artificial Neural Networks (ANN) generally performed
 735 better than multiple linear regressions (MLR), but MLR's still managed to explain a large proportion
 736 of the variance in photosynthesis.

737

	ANN			MLR	
	R _{eco}	GPP	NEE	R _{eco}	GPP
MAE (g C/m ² /day)	0.56	0.37	0.42	0.85	0.62
r ²	-	-	-	0.41	0.68
n	372	529	698	372	529

738

739

740

741 Table 3: Relative importance (percentage) of the different variables used to predict ecosystem

742 respiration, gross primary productivity, and net ecosystem exchange using an ANN.

743

R_{eco}		GPP		NEE	
f_{APAR}	36%	f_{APAR}	46%	f_{APAR}	27%
RAWC	19%	time since wetting	19%	RAWC	26%
PAR	18%	PAR	14%	time since wetting	14%
time since wetting	14%	RAWC	12%	water deficit	14%
water deficit	13%	water deficit	5%	T_{pn}	10%
T_{re}	0%	T_{pn}	4%	T_{re}	6%
				PAR	3%

744

745 Table 4. Results of a multiple linear regression to predict ecosystem respiration (a), and GPP (b). The
 746 best respiration model included f_{APAR} , time since wetting, soil temperature, and relative available
 747 water content, and two-way interactions between these variables. This corroborates the findings of the
 748 ANN model, but does not produce a good prediction ($r^2 = 0.41$, MAE = 0.85 (g C/m²/day)). The best
 749 GPP model included f_{APAR} , time since wetting, relative available water content, mean daytime
 750 temperature, and three-way interaction between several variables. This also corroborates ANN results,
 751 and produces a reasonable prediction ($r^2 = 0.68$, MAE = 0.62 (g C/m²/day))
 752 a)

	Std.				P
	Estimate	Error	t-value	P	
f_{APAR} :time since wetting	1.21	0.33	3.70	0.000	***
f_{APAR}	45.91	14.71	3.12	0.002	**
RAWC: T_{re}	0.02	0.01	2.92	0.004	**
time since wetting	-0.27	0.10	-2.80	0.005	**
f_{APAR} :PAR:time since wetting	-0.13	0.05	-2.56	0.011	*
f_{APAR} :time since wetting:RAWC	-0.03	0.01	-2.54	0.012	*
f_{APAR} : T_{re}	-1.48	0.62	-2.38	0.018	*
time since wetting:RAWC	0.01	0.00	2.33	0.020	*
f_{APAR} :RAWC	-0.36	0.18	-1.97	0.049	*
(Intercept)	-4.24	2.42	-1.75	0.081	.
RAWC	-0.19	0.13	-1.52	0.131	
T_{re}	0.16	0.11	1.49	0.139	
PAR:time since wetting	0.02	0.02	1.45	0.149	
PAR	0.22	0.28	0.78	0.437	
f_{APAR} :PAR	1.01	1.39	0.73	0.469	

753

754 b)

	Estimate	Std. Error	t-value	P	
RAWC	0.99	0.15	6.72	0.000	***
$f_{APAR:PAR:RAWC}$	0.40	0.08	5.35	0.000	***
$f_{APAR:RAWC}$	-1.89	0.42	-4.50	0.000	***
RAWC: T_{pn}	-0.02	0.01	-4.33	0.000	***
PAR:time since wetting:RAWC	0.00	0.00	-4.27	0.000	***
$f_{APAR:time\ since\ wetting:RAWC}$	0.03	0.01	4.25	0.000	***
PAR	2.00	0.52	3.88	0.000	***
$f_{APAR: water\ deficit}$	0.93	0.25	3.69	0.000	***
$f_{APAR:PAR}$	-6.33	1.75	-3.63	0.000	***
water deficit	-0.12	0.03	-3.50	0.001	***
PAR:time since wetting: T_{pn}	0.00	0.00	3.39	0.001	***
PAR:RAWC	-0.08	0.02	-3.29	0.001	**
$f_{APAR:PAR:time\ since\ wetting}$	0.17	0.06	3.08	0.002	**
PAR: T_{pn}	-0.05	0.02	-2.77	0.006	**
$f_{APAR:time\ since\ wetting: T_{pn}}$	-0.07	0.03	-2.73	0.007	**
PAR:time since wetting	-0.09	0.03	-2.68	0.008	**
time since wetting:RAWC	-0.02	0.01	-2.48	0.013	*
time since wetting:RAWC: T_{pn}	0.00	0.00	2.41	0.016	*
f_{APAR}	-31.56	13.95	-2.26	0.024	*
$f_{APAR:T_{pn}}$	1.06	0.61	1.74	0.083	.
(Intercept)	-5.34	3.27	-1.63	0.103	
T_{pn}	0.21	0.14	1.53	0.126	
$f_{APAR:time\ since\ wetting}$	0.77	0.63	1.22	0.223	
time since wetting	0.20	0.19	1.06	0.291	
time since wetting: T_{pn}	0.00	0.01	-0.26	0.792	

755 Table 5: Summary of NEE over the 5 year period for which there was flux data. Negative values
 756 represent an overall sink of carbon. Data gaps were filled using an ANN and predictors f_{APAR} , water
 757 deficit, relative soil moisture content, mean day time temperature, time since wetting, and mean soil
 758 temperature, in that order of importance.
 759 Also reported are annual summaries of rainfall, available photosynthetically active radiation, length of
 760 the growing season, and number of growth days (days when soil moisture content is greater than θ_{crit}
 761 (7% by volume).

762

Rainfall year (July to June)	Annual NEE (gC/m ²)	95% confidence interval	Annual rainfall (mm)	Annual PAR (MJ/m ²)	Growing season length (days)	Number of growth days
00_01	42	(17; 67)	659	662	244	245
01_02	155	(130; 180)	572	523	191	169
02_03	150	(125; 175)	303	406	156	166
03_04	-138	(-163; -113)	618	555	188	81
04_05	-83	(-108; -58)	760	665	197	186

763

764

765

766

767 Table 6: Annualised summary of the different contributions to the carbon balance at the Skukuza flux
768 site.
769

	Mean annual flux
Herbivory	9.5 g Cm ⁻² y ⁻¹ (unknown error ? 20%)
Fire	33.6 ± 14.7 g Cm ⁻² y ⁻¹
Flux measurement (incl. herbivory)	75 ± 105 g Cm ⁻² y ⁻¹
Total	108.6 ± 119.7 g Cm ⁻² y ⁻¹

770

771 **10. Appendix A: Comparison of meteorological data**

772 Correlation between the flux tower variables and corresponding variables from other sources
773 appears in table A1.1. Strong linear relationships exist between the flux tower daily
774 measurements for the mean soil temperature and the mean daytime temperature and the
775 corresponding temperature variables derived from the minimum and maximum daily
776 temperatures of the South African Weather Services (SAWS) data. There is also a strong
777 linear relationship between the measured mean soil moisture and the modelled soil moisture
778 using the SAWS data, as well as a fairly strong linear relationship between PAR derived from
779 the shortwave radiation from the flux tower and the modelled PAR.

780

781 The correlation between the flux tower rainfall and the SAWS rainfall is significant, but not
782 as strong as that of the previous comparisons to SAWS derived variables. The peaks of the
783 environmental data are usually slightly higher than recorded from the flux tower, although
784 there are few days when the flux tower recorded higher values (fig A1.2). This could be due
785 to localised rainfall events. Peaks in the data do not always correspond and this could be due
786 to the measurements from the SAWS data being taken daily from a rain gauge, whereas the
787 flux tower took instantaneous measurements of rainfall. Therefore daily rainfall events may
788 not always correspond exactly. The pattern of rainfall during time appears to match for the
789 two data sets. The annual sum of rainfall for the environmental data is always more than that
790 for the flux tower data (Table A1.1). This is due to missing data from the flux tower.

791

792 There is a strong linear relationship between Gimms NDVI and f_{APAR} (Table A1.2).

793 Therefore a linear regression equation was derived to describe this relationship. The linear

794 regression obtained a R^2 -value of 0.7072 and an MAE of 0.0520. The estimated equation

795 was: $f_{APAR} = -0.079 + 0.736 \times \text{Gimms}$

796 The standard error for the intercept is 0.004 and the standard error for the slope is 0.009.

797

798 Table A1 Summary of comparisons between flux tower derived variables and corresponding

799 variables derived from other sources

Variables compared	Pearson correlation	95% confidence interval
Mean flux tower soil temperature and derived soil temperature from SAWS data (T_{re}).	0.9242	(0.9154; 0.9322)
Mean flux tower daytime temperature and derived daytime temperature from SAWS data (T_{pm}).	0.9558	(0.9507; 0.9603)
Scaled flux tower soil moisture and derived scaled soil moisture from SAWS data (θ_{rel})	0.7752	(0.7500; 0.7981)
Daily flux tower rainfall and SAWS rainfall data.	0.6109	(0.5824; 0.6378)
f_{APAR} and GIMMS NDVI.	0.8409	(0.8294; 0.8517)
PAR calculated from the flux tower data and the modelled PAR data	0.6237	(0.5841; 0.6604)

800

801 Table A2: Annual rainfall over time

<u>Annual Rainfall Sum from SAWS Environmental Data</u>						
99/00	00/01	01/02	02/03	03/04	04/05	05/06
363	659	572	302	618	760	249
<u>Annual Rainfall Sum from Flux Tower Data</u>						
99/00	00/01	01/02	02/03	03/04	04/05	05/06
415	671	427	310	276	582	209

802

803 **11. Appendix B: Interpolating day-time respiration**

804 Fitting an optimal temperature function to the mass of night-time flux measurements involved
805 making several assumptions about a) the shape of the temperature-respiration curve, and b)
806 the values to use to fit the curve.

807

808 ***11.1. Shape of the temperature-response curve***

809 Field data indicate that a generalised Poisson function is the best descriptor of the effect of
810 temperature on respiration, as it describes both the exponential increase of respiration with
811 temperature and the sudden decrease once the temperature optimum has been reached (Kirton
812 et al in prep). However, for this analysis we also tried a simple parabolic function.

813 ***11.2. Values used to fit the curve***

814 This interpolation method relies on deriving a curve that represents the temperature response
815 under a certain set of environmental conditions. Any deviation from this line by an observed
816 point is then assumed to be due to different environmental conditions. The curve can be
817 pulled up and down to match this point, and thereby adjust for these varying environmental
818 conditions, by the use of a scaling parameter. Missing respiration values (day time points) can
819 then be interpolated on this day (because the environmental conditions other than temperature
820 are going to remain stable at a daily time step) by using the temperature at each point and the
821 adjusted temp/resp equation.

822

823 With this in mind, extracting the points to be used could be done in a number of different
824 ways. The easiest way to identify points where all factors other than temperature are constant
825 would be to identify the maximum points for each temperature value (which would represent

826 respiration under completely optimal conditions of soil moisture and LAI). We tried three
827 different methods for extracting these values: manually picking the maximum respiration
828 values, calculating the maximum respiration value for each degree temperature change, and
829 calculating the 95th quantile for each degree temperature change (Figure B1). We also tried
830 manually picking values at the top of the thickest part of the cloud of respiration points. This
831 approach would exclude any extreme outliers but could also be assumed to represent the
832 same set of other environmental conditions. Because the curve is adjusted up and down based
833 on the respiration values on the day in question, the position of the curve on the y axis is
834 unimportant. It is the shape of the curve that will affect the interpolation.

835

836 Using the 95th quantile was not satisfactory as some temperature categories had orders of
837 magnitude more respiration measurements than others. We therefore abandoned that method
838 and tested six different respiration interpolation methods (Table B1): manually selected
839 maximum points (fitting a parabolic and generalised Poisson), manually selected points at
840 edge of data cloud (parabolic and GDP), and calculated maximum points (parabolic and
841 GDP).

842 **11.3. Results**

843 Results indicate that the interpolated values are very resilient to the method used to fit the
844 temperature response curve. The distribution of interpolated points was similar for all six
845 methods (Figure 2), and linear regression models show similar fits to the observed respiration
846 data (Table 1). A visual assessment of the interpolated points (Figure 3) indicates that the
847 generalised Poisson interpolations fell more clearly within the main data cloud. We therefore
848 chose to use the calculated maximum value method fitted to the generalised Poisson
849 distribution.

850

851

852 Table B1: The six different methods used to fit a temperature response curve to the measured
 853 night-time (respiration) fluxes. Two different fitting functions were used, and three different
 854 methods for identifying points to fit the curve to. The distributions of the data interpolated
 855 with each method were very similar to each other (Figure 2), and fell well within the bounds
 856 of the observed respiration data (Figure 3).

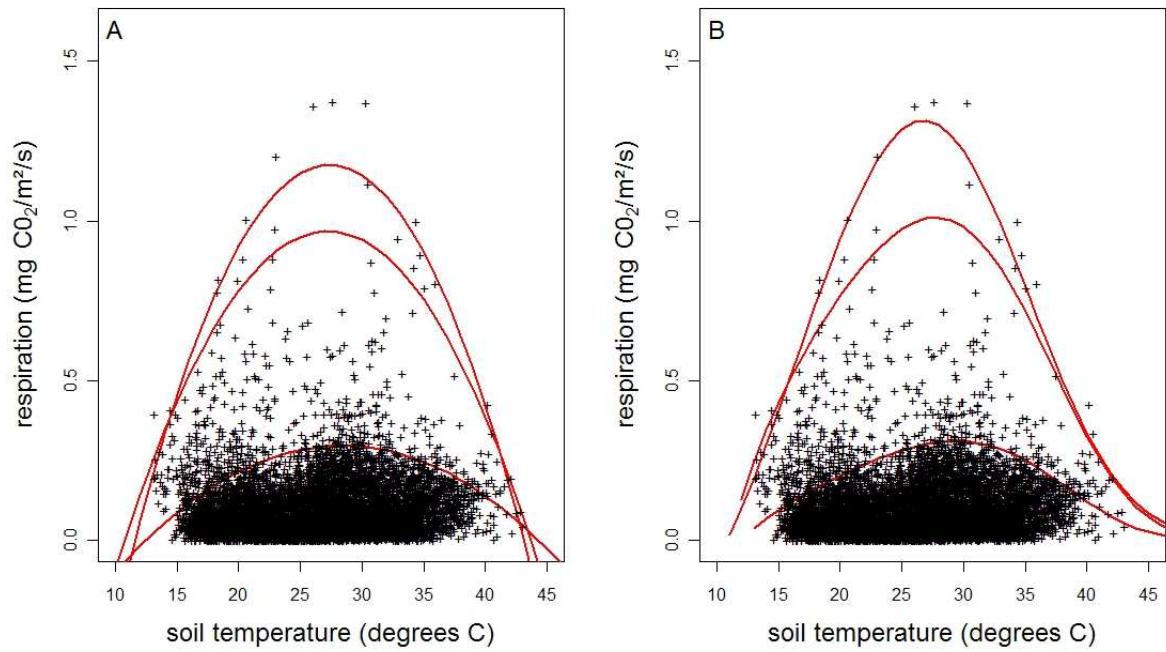
	Parabolic			Generalised Poisson		
	Observed max	Calculated max	Observed fit to datacloud	Observed max	Calculated max	Observed fit to datacloud
Name	parObsMax	parCalcMax	parObsMain	poisObsMax	poisCalcMax	poisObsMain
r ²	0.57	0.58	0.56	0.56	0.58	0.56
slope of linear model	0.61	0.61	0.6	0.56	0.58	0.56
Median predicted value						
mgCO ₂ /m ² /s	0.070	0.070	0.069	0.071	0.070	0.069
Minimum predicted value						
mgCO ₂ /m ² /s	0.002	0.002	0.000	0.002	0.002	0.000
Maximum predicted value						
mgCO ₂ /m ² /s	0.98	0.68	0.81	0.81	0.71	0.81

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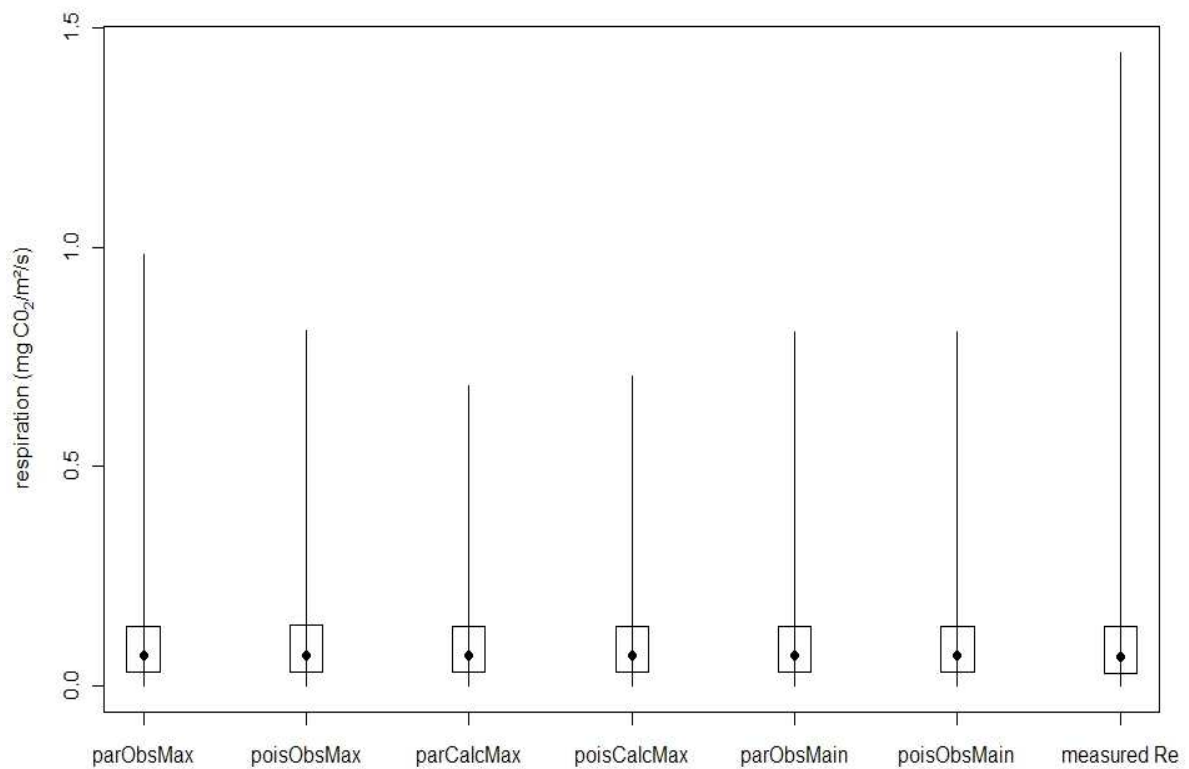
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862

863 Figure B1: Showing the six temperature response functions fitted to the half-hourly night
864 time fluxes (respiration). Plot A shows the parabolic functions fitted over the manually
865 selected maximum points (top function), the automatically selected maximum points (middle
866 function) and the manually selected top of the data mass (bottom function). Plot B shows the
867 Generalised Poisson function fitted over the same three selections of points.

868

869



870

871

872 Figure B2: Showing the distribution of the respiration data interpolated using six different
 873 methods (solid points: median values, box: +/- 25 % quantiles, bar: data range). The median
 874 and +/- 25 % quantiles are very similar for each method, but the method that calculates the
 875 fitted values had slightly lower maxima than the other two methods. All data are well within
 876 the range of measured Re values (u*-corrected half-hourly night-time fluxes).

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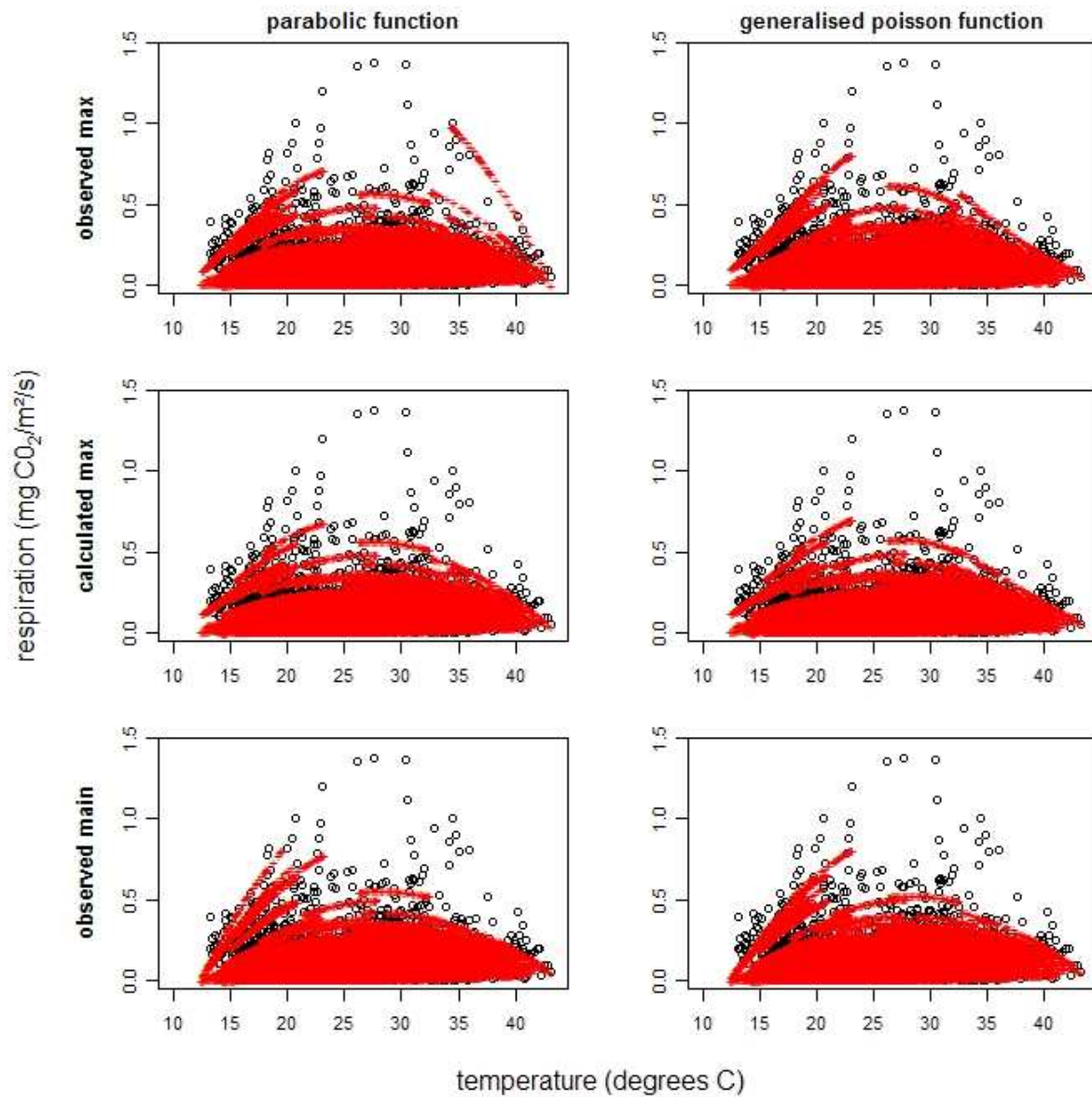
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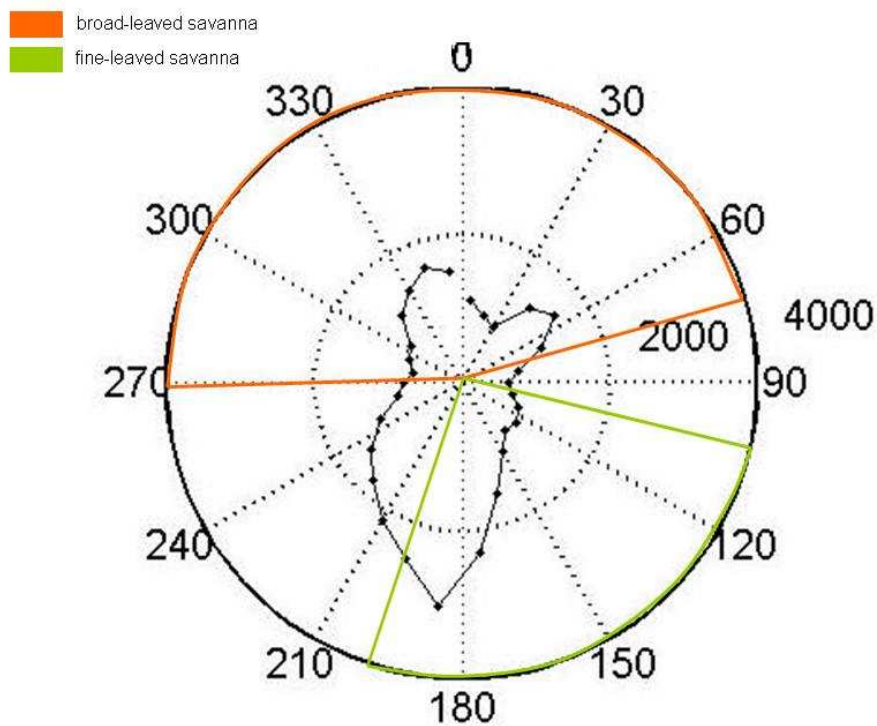
888 Figure B3: The distribution of measured half-hourly night-time fluxes (black circles) and
 889 interpolated half-hourly respiration (red crosses) along a temperature axis. Interpolated fluxes
 890 represent all half-hour values which had soil temperature data and at least three night-time
 891 fluxes to estimate the scaling parameter.

892 **12. Figures referred to in the response to reviewers**

893

894

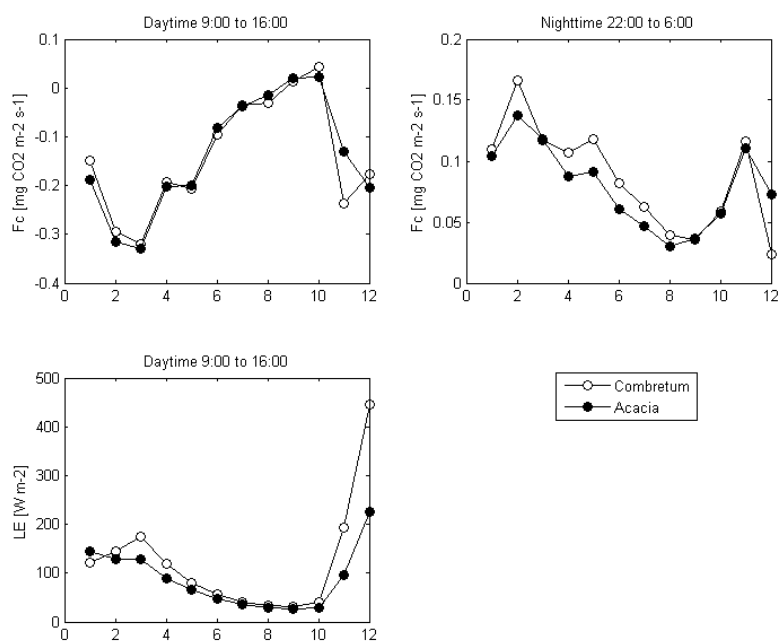
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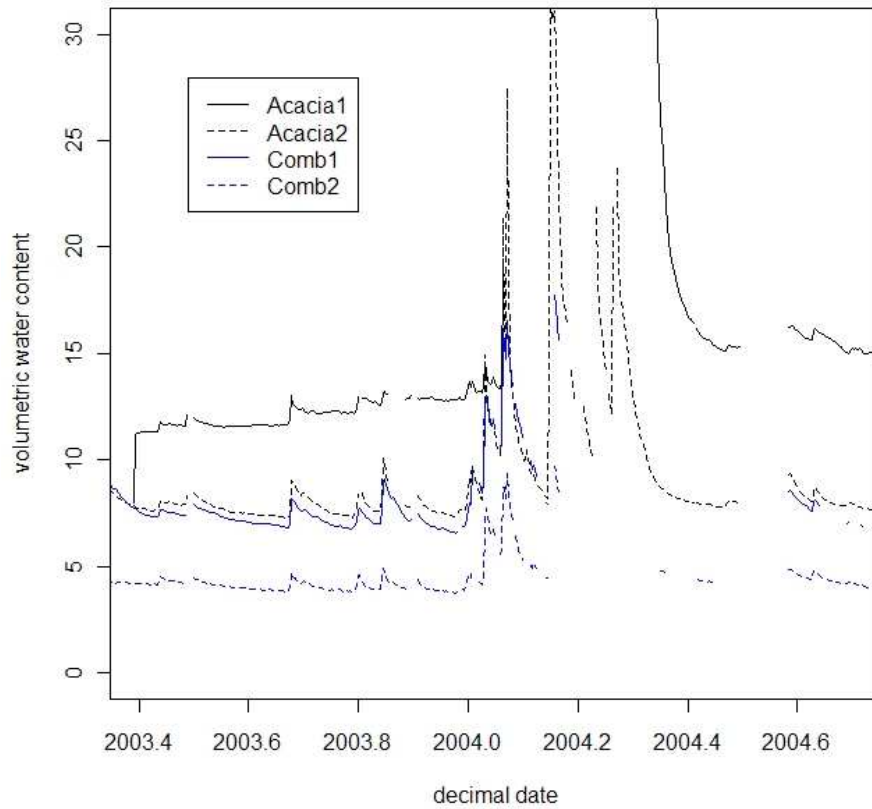
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898 Figure 1: Showing dominant wind directions at the flux site, and the approximate fetch of the
 899 broad-leaved *Combretum* savanna and fine leaved *Acacia* savanna
 900

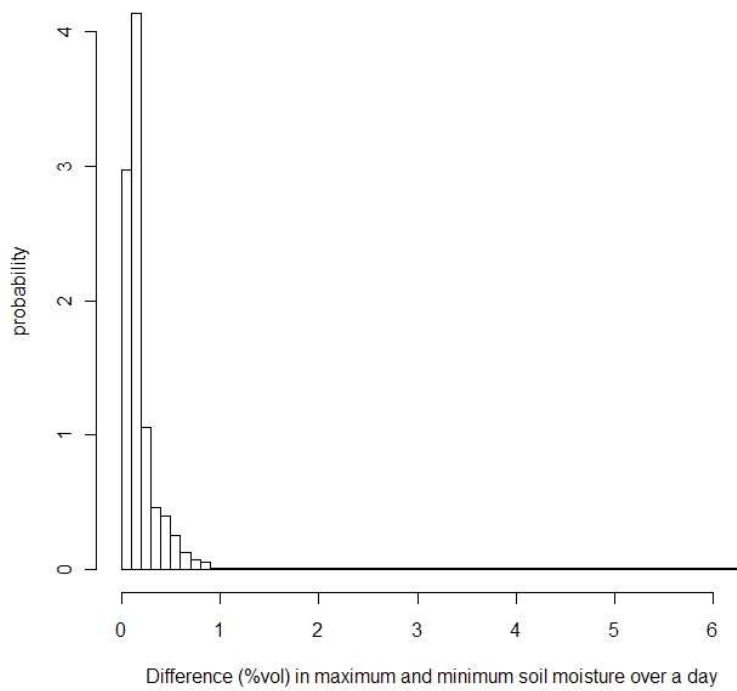


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902 Figure 2: Showing mean monthly carbon (Fc) and water (LE) flux at the site for the two main
903 vegetation types (calculated from the wind directions). Carbon fluxes are separated into
904 daytime and nighttime fluxes.
905



906
907
908 Figure 3: Soil moisture time series from the four sets of soil moisture probes at the site. Data
909 presented as volumetric water content over the entire profile.
910
911
912



913
 914 Figure 4: Probability distribution of the range in soil moisture values over one day (midnight
 915 to midnight) at the flux tower. The soil moisture difference was less than 0.2% in about 75%
 916 of the days.
 917